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R. B. Hunter

A. Wallace

E. M. Romney

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FINAL REPORT

NITROGEN TRANSFORMATIONS IN ROCK VALLEY AND ADJACENT AREAS OF THE MOHAVE DESERT

R. B. Hunter, A. Wallace, and E. M. Romney University of California, Los Angeles

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Ecology Center, Utah State University, Logan, Utah 84322

ABSTRACT

This is the final report of investigations of nitrogen transformations in the northern Mohave Desert areas adjacent to Rock Valley. Harvests and tissue analyses of *Ambrosia dumosa* showed that within measurement limits all nitrogen required for growth of leaves and new stems during 1975 was derived from soil pools. Although nitrogen in leaves ranged from O to 40 % of **N** in live above-ground tissues, it was not depleted from stems or large roots in spring, nor accumulated in them as dormancy approached. Variations in *Ambrosia* leaf and stem concentrations of other elements during 1975 were also measured. A buildup of soluble salts in leaves suggested that they act as wicks after capillary flow of water to roots ceased. Nitrogen in 15N-tagged plant material added to Rock Valley soils was more available to plants than endogenous soil nitrogen. A half-life of three to four years for this added nitrogen was indicated by soil analysis and plant uptake. Loss solely through harvest would result in a half-life of seven to eight years under glasshouse conditions. Treatment of selected shrub clumps with 15N fertilizer allowed estimates of root zone radii varying from less than one to more than 7 m for common Mohave Desert shrubs. Labeling of plant tissues occurred primarily in new leaves, stems and fruit, and all soil analyses showed 15NO; remained in the nitrate form 70 days after application. Extensive measurements of acetylene reduction indicated a near absence of nitrogen fixation associated with random samples of soils, litter or roots of major shrub species. **A** seasonal trend of ethylene production by roots was found. Evolution of ethylene from litter and adsorption of ethylene and acetylene by dry soils were indicated. Soils amended with fertilizers and incubated under the field under three watering regimes showed ammonia volatilization, nitrification, denitrification and organic matter decomposition to be slow processes under Mohave Desert conditions. Input of ammonium and nitrate by rainfall amounted to approximately 2 kg N/ha in 1976. Three samples of cemented gravel from the undersides of buried rocks contained 306, 398 and 621 *µ* g NO₃ N/g of acid-soluble "cement."

INTRODUCTION

Nitrogen is the most commonly limiting plant nutrient and plays an important role in uptake of other essential elements (Hiatt and Legget 1974; Raven and Smith 1976). Although water is more commonly limiting in the Mohave Desert, nitrogen fertilization has been shown to increase yield and N content of both annuals and shrubs (Romney et al. 1974; Hunter et al. 1975a, 1976a). The size of nitrogen pools in the Mohave Desert, additions to them from the atmosphere and transformations between pools have been investigated under this project for the last four years (Wallace et al. 1974b; Hunter et al. 1975b, 1976b), following some previous work under other fundings (Wallace et al. 1972; Romney et al. 1973) and by other researchers (Wells 1967; Garcia-Moya and McKell 1970). It has become increasingly clear that nitrogen dynamics in the Mohave are different from those in the cool deserts (Rychert and Skujins 1974; Reichle 1975), especially in the relative importance of lichen and algae crusts.

In this final report we present results relating to plant storage of nitrogen, mineralization of plant litter and soil organic matter, denitrification, nitrogen fixation by lichen-algae crust, rhizosphere microorganisms and decomposing microorganisms, input of nitrogen in rainfall, volatilization losses of soil ammonia and the half-life of N in a natural shrub community.

METHODS

Ambrosia MINERAL TURNOVER

Three small-to-medium *Ambrosia dumosa* plants were harvested at intervals throughout 1975. They were separated into tissues and analyzed for N and minerals by Kjeldahl digestion and emission spectrography (Wallace et al. 1974a), respectively.

LrrrER DECOMPOSITION

Ground samples of "N-tagged desert plant species were added to pots of Rock Valley soil (from several locations, sieved and mixed) in January 1975. Crops of native annuals have been grown in these pots several times a year since. Nitrogen-15 enrichment was measured by emission spectrography directly in plant tissues, and in soils after Kjeldahl extraction and precipitation as $(NH_4)_2$ PtCl₆.

UPTAKE OF ¹⁵N FERTILIZERS

Application of $^{15}NO_3^-$ and $^{15}NH_4^+$ fertilizer to four 5-m² areas was described previously (Hunter et al. 1976b). Analyses were by emission spectrography as described above.

N-BALANCE IN FIELD-INCUBATED SOILS

One hundred kg of two different soils was collected in Mercury Valley, sieved (2 mm) and each thoroughly mixed in a cement mixer. One was from a large bare area covered by desert pavement with sparse shrub cover. The other was taken from beneath a shrub clump dominated by a large *Ephedra nevadensis* (1 m height, 1.2 m diameter). In this area *Ephedra* builds up a fairly thick litter mat. Before sampling, a surface sample from each site showed nitrate-N concentrations of ≤ 2.5 μ g/g in the bare area and $\sim 20 \mu$ g/g under the shrub. Samples were taken to a depth of 30 cm.

These soils were treated in various ways (described below) and divided into 774 samples in order to determine rates of ammonia volatilization, denitrification, nitrification and ammonification.

To estimate ammonia volatilization, 8 .4 kg of bare soil was steam sterilized at 15 lb for 40 min. N-Serve (2-chloro-6 [trichloromethyl] pyridine, Dow Chemical Co. lot WP10294-2) was diluted .25 ml/836 ml with distilled water,

added to and thoroughly mixed with the soil. The soil was dried for three days at 60 C in a forced draft oven. Finally 0.987 g (NH₄)₂SO₄ was ground to a fine powder and thoroughly mixed with half the soil.

For determination of litter N mineralization, the bare sample was amended with ground litter collected from Rock Valley. The litter included material from beneath six common shrub species, and was added at the rate of 3 g/100 gdw soil. Unamended shrub and bare soils were also used.

For the nitrification study, 0.978 g of powdered (NH_4) ₂SO₄ was added to 2.1 kg bare and shrub soils.

To determine denitrification, N-Serve was added as previously described. Nitrate was added by dissolving 3.53 g $Ca(NO₃)₂ · 4H₂O$ in water, which was then mixed with 4.2 kg bare and shrub soil.

Soils were apportioned into 774 74-ml acrylic plastic vials and set outside under a Plexiglas cover (with sides open to wind). One-third were not watered, one-third received "normal" water and one-third received two times "normal" water. Rainfall was averaged by amount and number of events for each month to determine amount and timing of watering. Actual water applied, given in Table 1, differs slightly from theoretical timing due to vacations and weekends. Prior to addition of water, 10 vials were weighed to prevent overwatering. Moisture held by soils in three thoroughly wet vials allowed to drain 48 hr averaged 25.8 \pm 0.3 g (se). Field-incubated vials never approached this quantity of water, indicating average rainfall conditions in the Mohave are not suitable for thorough wetting of the soil.

Twenty-four vials of bare soil with added litter were wet with 25.5 g water and incubated in a lab cupboard ($\sim 21 \text{ C}$) and refrigerator (\sim 5 C). A small piece of tape was applied to the rim to prevent total sealing by the plastic caps. No water was applied throughout the year, as incubation conditions did not allow an approximation of field moisture conditions. These soils were still thoroughly wet at the end of the one-yr incubation.

Three vials from each treatment were removed immediately after placement in the field April 7, 1976, and also one week, one, three and six months, and one year after the start of incubation. The ammonia volatilization experiment was duplicated starting July 7 with a new bare soil, as rainwater dripping from the supporting structure for the Plexiglas cover flooded a number of vials from the experiment.

Analyses for NO_3 , exchangeable NH_4^+ and total N were made with selective ion electrodes (Orion). Nitrate was extracted with 0.01 M sodium citrate (2:1, V/soil dw). Ammonium was extracted with 2N KC! (10:1) and total N by Kjeldahl digestion.

ACETYLENE REDUCTION

The gas chromatograph was moved from UCLA to the CETO lab in Mercury for work done in 1976, hence samples

Table 1. Water added to N-balance soils, field-incubated in 74-ml vials

were run within hours of removal from the field. For root samples, serum bottles were incubated in the field and buried in the hole from which they were taken. Soils were incubated in the CETO glasshouse, and microorganisms in 35-ml bottles in the lab.

Nitrogen-free agar used for culturing litter isolates contained in g/liter, 1.0 **K,HPO,,** 0.2 **MgSO,,** .01 NaCl, .01 FeSO,, .01 MnSO,, 20.0 glucose, 30.0 CaCO, and 0.41 **KOH.** Ten ml of hot agar was added per 35-ml bottle before sealing and sterilization (15 lb, 30 min). Antibiotics and inoculant were injected through the serum bottle cap after sterilization.

NITRATE **AND AMMONIUM IN RAIN**

A standard funnel-type rain gauge with the funnel neck lightly plugged with glass wool was used to collect rainfall for nitrogen analysis. Nitrate and ammonium content were measured when enough sample was available using Orion selective ion electrodes.

CALICHE NITRATE

Three samples of cemented gravel and deposits on the underside of large rocks were partially dissolved in 50% H,SO,. Nitrate was determined by the known addition method using an Orion nitrate electrode.

RESULTS

Ambrosia **MINERAL** TuRNOVER

Allocation of nitrogen to leaves and stems of *Ambrosia dumosa,* adjusted to 100 g live shoot weight, is shown in Figure 1. Over a period of 70 days, leaf N increased from 4 to 40 % of total N in the live above-ground parts. As the season progressed, new stem growth accounted for up to 13 % of

| Date | Live Stem | Leaves | New Stem $\frac{d}{p}$ + se. | Large Root | Fine Root |
|------------------|---------------|---------------|---------------------------------|-----------------|---------------|
| Jan 17 | $0.83 + 0.09$ | | | $1.27 + 0.07$ | $1.26 + 0.23$ |
| Feb 3 | $0.67 + 0.08$ | ä, | | $1.03 + 0.21$ | $1.34 + 0.13$ |
| Feb 18 | $0.82 + C.08$ | ú, | Sec. | $1.22 + 0.10$ | $1.41 + 0.05$ |
| Mar ₃ | $0.80 + 0.10$ | 1.85 | ÷. | $1.02 + 0.11$ | $1.27 + 0.22$ |
| Mar 17 | $0.84 + 0.11$ | $5.85 + 0.29$ | \cdot | $1.17 + 0.05$ | 1.10 |
| Apr 1 | $0.83 + 0.08$ | $4.28 + 0.17$ | \sim | $1.03 + 0.05$ | $1.32 + 0.02$ |
| Apr 14 | $0.96 + 0.07$ | $4.59 + 0.25$ | $\overline{}$ | $2.11 + 0.09$ | $1.33 + 0.33$ |
| Apr 28 | $0.75 + 0.04$ | $4.14 + 0.49$ | $3.12 + 0.52$ | $0.93 + 0.16$ | $1.18 + 0.20$ |
| May 12 | $0.76 - 0.04$ | $3.38 + 0.09$ | $1.95 + 0.08$ | $1.15 + 0.21$ | |
| May 27 | $0.66 + 0.02$ | $2.69 + 0.29$ | $1.18 + 0.04$ | $0.74 + 0.05$ | $1.06 + 0.05$ |
| Jun 12 | $0.70 + 0.05$ | $2.08 + 0.25$ | $0.96 + 0.05$ | 0.92 ± 0.04 | $0.74 + 0.16$ |
| Jun 30 | $0.82 + 0.07$ | $1.75 + 0.04$ | $1.07 + 0.07$ | $1.21 + 0.03$ | 1.01 |
| Jul 21 | $0.89 - 0.15$ | $1.30 - C.05$ | $1.02 + 0.19$ | $1.20 + 0.11$ | .26 |
| Aug 18 | $0.70 + 0.05$ | $1.59 + 0.37$ | $0.93 + 0.11$ | $0.8y + 0.15$ | |
| Sep 17 | $0.81 + 0.01$ | $1.17 + 0.05$ | $0.95 + 0.04$ | | $1.33 + 0.84$ |
| Oct 21 | $0.73 - 0.08$ | 1.23 | | 0.99. | |
| Nov 20 | $0.81 + 0.10$ | $1.6 + 0.69$ | 2.57 | $0.63 + 0.22$ | |
| Dec 16 | $0.62 + 0.02$ | 1.18 | \sim | $0.90 + 0.10$ | $0.75 + 0.1$ |

Table 2. Concentration of N in live tissues of Ambrosia dumosa during 1975

total N. Gain and loss of N in the shoot generally parallels growth and loss of leaf dry weight (see Hunter et al. 1976b; Fig. 1).

Though N concentration from plant to plant is quite variable, there is no indication of a significant decrease in stem N during the spring leaf flush, nor of an increase as the plants become dormant. Hence, stems do not act as a storage site for nitrogen.

Changes in N concentration of leaves, live stem, new stem, large roots and fine roots are shown in Table 2. When first large enough to sample, new leaves contained four to seven percent N, while young stems contained two to four percent. When dormant in September, leaves contained about 1.2% and stems near 1 % N. At this time most of the leaf surface was dead, with just the midribs and youngest leaves being green. Dormancy was not sudden, as with deciduous trees, but rather occurred as a slow death of leaves, from older to younger and from outer edge to the midribs. Ambrosia leaves do not abscise as many species, but remain attached to the stem for several months after dying. Figure 1 and Table 2 do not differentiate live from dead leaves because of this pattern.

There are significant ($p = .05$) negative correlations of N concentration with time in new stem, leaves, large and fine roots, and dead stem (Tables 2 and 3). Large root tissue includes the crown, which after stem tissue is the only likely place for nitrogen storage during the dormant period. Hence we find no evidence that Ambrosia stores nitrogen, implying N for the spring growth flush is taken up from the soil.

Figure 2 shows changes in the ratios of leaf concentration to live shoot concentration of several elements. For leaf N the decrease to late May (day 147) can be largely explained by carbohydrate dilution, but after that date it indicates a minor "withdrawal" of N from the leaves. As noted above, however, we have no indication the withdrawn N is stored in any given Auzzit

The variability in N concentration in old live stem is partially explained by percentage of dead wood. There is a positive correlation ($r = +0.33$, $p < 0.025$) of percent dead wood to stem N, and also a significant correlation ($r = -.52$, $p = .05$) between the ratio of live to dead stem weights and the leaf: live stem ratios between April 28 and June 30 (i.e. while fully leafed out). These data suggest the live stem competes with leaves for nutrients, and that death of some stems allowed for increased leaf:stem ratios and stem N concentrations in remaining branches.

Figure 1. Above-ground nitrogen fractions in Ambrosia dumosa during 1975 in Mercury, Nevada.

Figure 2. Ratios of leaf concentrations to average concentrations in live above-ground portions of Ambrosia dumosa during 1975 for several elements. Error bars are 95% confidence limits, 1-3 samples at each date

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| | Stem | $\frac{\text{New}}{\text{Stem}}$ | Dead | Leaves | Dead Leaves | Large Root | Fine Root |
|--|---------|----------------------------------|-------------------------|---------|----------------|---------------|--------------|
| $ r > 0$, $p = .05^a$ | .29 | | $rac{\text{Stem}}{.28}$ | .34 | .50 | .27 | .29 |
| | | | | | | | |
| Element | | | | | | | |
| N P Na K Ca K Ce K Fe Zn | | $-.69$ | $-.29$ | $-.79$ | | $-.54$ | $-.53$ |
| | $-.46$ | | $-.33$ | $+, 34$ | | $-.28$ | $-.38$ |
| | $+, 49$ | $+, 58$ | $+, 33$ | $+.77$ | | $+, 42$ | |
| | | | | $-.41$ | $+, 86$ | | |
| | $-.38$ | | | $+, 51$ | | | $-.31$ |
| | | | | $+, 69$ | $+.64$ | | $-.37$ |
| | $-.37$ | | $-.40$ | | | $-.48$ | $-.37$ |
| | $-.36$ | | $-.53$ | $+, 34$ | | | |
| Cu | $-.33$ | | $-.32$ | | $-.61$ | | |
| | | | | $+, 50$ | | | |
| $\begin{array}{l} \mathtt{M} \mathtt{n} \\ \mathtt{B} \\ \mathtt{S} \mathtt{i} \\ \mathtt{S} \mathtt{i} \end{array}$ | | $-0.5h$ | | $+, 57$ | $+, 74$ | | |
| | $-.43$ | | | $+, 55$ | | | $-.31$ |
| | | | | $+.60$ | | | |

Table 3. Tissue mineral concentrations in *Ambrosia dumosa* correlating significantly with time between January and December, 1976

 4 Significant $|r|$ was determined using the relationship:

F (1, n - 2) , $95 = r^2$ (n - 2)/(1 - r^2),

Soil nitrate concentrations under the harvested shrub ranged from 0.1 ± 0.1 to 6.2 ± 1.4 ppm (avg \pm se, n = 3) per sample period) for the 17 sampling dates. There was no seasonal pattern, and no correlation of soil N with stem N. These low concentrations are typical of the open areas where these plants were harvested. (Plants growing in clumps were not harvested.)

DSCODE A3UTJ3 contains tissue analyses for 12 elements other than N. They areP, Na, **K,** Ca, Mg, Zn, Cu, Fe, Mn, B, Al and Si. Figure 2b, c and d shows variations in ratios of leaf concentration to average shoot concentrations for Ca, **K** and P. It appears that as leaves die (after day 181) there is a concentration of Ca and Pin them, with some indication of a late drop ih **K.**

In 1975 there was very little rainfall in Mercury after May 21, and soil water potentials were lower than -30 bars at 15. 30 and 60 cm from June through December (Fig. 3e). During that period there would have been negligible capillary flow of water to roots, and essentially no uptake of mineral ions from soil. It appears from the correlations given in Table 3 that most elements moved from large roots and stems into leaves with time, which may indicate leaves act as a wick, collecting salts while they play a major role in transpiration. Sodium, however, did not fit this pattern, increasing in concentration in all tissues with time. The particularly significant increase in leaf and dead leaf concentrations of Na, B, Mg and **K** (dead leaves only) suggest the wick effect. Increases in Si, Al and Mn are often due to dust contamination, but may also be due to uptake of these elements through roots and their concentration in dying leaf tissue with the above salts.

There is a high correlation between Si and Al in leaves $(r =$.96), but not in stems $(r = .29)$. We believe the best explanation for this involves dust contamination of leaves. Data of Romney et al. (1974) bear on this question. In stems there is a strong correlation between Fe and Al $(r = .89)$, which may indicate similarity in uptake mechanisms, or a single uptake process for trivalent cations.

Figure 3 plots stem concentrations of several elements and water potential against time. We consider stem tissue to be least affected by developmental changes and to be the most stable in its mineral relations. Minerals changing significantly over the year are listed in Table 3. It appears from Figure 3 that several changes occur near mid-year, when soil water capillary flow ceased. In particular there was a loss of stem P and Ca and a gradual increase in stem Na after June. The strong correlation between Al and Fe concentrations is apparent in Figure 3.

LITTER DECOMPOSITION

Harvest of plant tissue and **N** from pots amended with ground 15N-tagged plant material is reported in Table 4. Both dry matter and nitrogen harvested were considerably enhanced by the addition of all types of litter. There are significant correlations between weight and N harvested and **N** added, and weaker correlations between those yields and dry weight added. In several cases the weight of plant material harvested to January 7, 1977 exceeded the weight added to soils, but **N** harvested has not exceeded two-thirds of that added. Yields in amended soils are two to three times control yields.

The availability of **N** from ground litter is greater than that of endogenous soil nitrogen (Table 5). Although the amendments made up 20 to 60 percent of total nitrogen, they contributed 56 to 99 percent of initial harvest **N** (excepting *Larrea* leaves), and two years later still generally provided a disproportionately high percentage of harvest N. This indicates endogenous soil **N** is considerably reduced in availability in comparison to litter N. This is reflected in the A-values reported earlier (Hunter et al. 1976b) of 55 to 218 mg **N** /kg soil, whereas Kjeldahl digests average initial **N** values of 450 mg **N** /kg. Hence it appears just 10-50 % of Rock Valley soil nitrogen was available over the *6* months during which A-values were determined.

Figure 3. Concentration fluctuations of several elements in old live stem of Ambrosia dumosa and maximum soil water potentials measured at 15, 30 or 60 cm during 1975. Error bars are 95% confidence limits, $n = 1-3$.

Table 4. Nitrogen and dry weight harvested from Rock Valley soils treated with ground tissues of several desert plant species. Harvested weight and nitrogen correlate significantly at $p = .05$ with nitrogen added ($r = .58$ and .72, respectively). They correlate less well with dry weight added $(r = .49$ and .58, the latter significant at .05)

| | | | Harvest To Jan. 7, 1977 | |
|----------------------|---------------|--------------------------------------|-------------------------|-------|
| Species and Tissue | Dry Wt, g | N. mg | Dry Wt, g | N, mg |
| Atriplex hymenelytra | | | | |
| root | 54.0 | 912 | 19.5 | 279 |
| stem | 55.5 | 732 | 17.8 | 212 |
| leaves | 77.8 | 2121 | 14.8 | 354 |
| Lycium andersonii | | | | |
| root | 81.3 | 1950 | 26.6 | 347 |
| stem | 109.4 | 2345 | 28.6 | 497 |
| leaves | 17.9 | 802 | 27.9 | 519 |
| Larrea tridentata | | | | |
| root | 27.6 | 547 | 15.2 | 205 |
| stem | 35.9 | 708 | 25.5 | 288 |
| leaves | 27.8 | 655 | 23.1 | 262 |
| Ambrosia dumosa | | | | |
| root | 19.5 | 386 | 20.0 | 256 |
| stem | 28.4 | 452 | 21.0 | 260 |
| leaves. | 29.2 | 1180 | 31.5 | 531 |
| Controls | | | | |
| | | 0 | S.0 | 96 |
| $\frac{1}{2}$ | $\frac{0}{0}$ | | 9.2 | 103 |
| | | $\begin{matrix} 0 \\ 0 \end{matrix}$ | 8.5 | 90 |

Soil analyses were erratic, due to the involved analytical procedures. Nevertheless, the loss of ¹⁵N indicated by soil analysis was greater than can be accounted for by harvested ¹⁵N (Table 6). Loss of added N was greater than loss of endogenous N. These soil data would indicate loss by denitrification or ammonia volatilization. It is supported somewhat by the decrease in labeled N as a percent of harvest N (Table 7). This decrease is on the order of that indicated by the soil analysis, with average half-lives of 3.3 yr for soil ¹⁵N decrease and 3.7 yr for decrease in ¹⁵N uptake. In contrast, if all N had been lost by harvest the average half-life would be 7.5 yr. Perhaps we should note here that the half-life discussed is for residence time and plant availability in soil, as ¹⁵N is not radioactive.

The decomposition of all tissues other than Larrea leaves appears to be similar, with half-lives (measured by plant uptake) of 1.3 to 5.7 yr. Larrea leaves show a much longer half-life of 38 yr, in agreement with the low initial N availability noted last year. The 5.7-yr half-life of nitrogen in Atriplex hymenelytra leaves also suggests delayed availability. It was very difficult to establish plants in the pot with ground A. hymenelytra leaves. Mineral analysis of Bromus rubens tissue from the harvest of April 28, 1975 showed the plants growing in this soil to be higher in Na, K, Zn, Cu, Sr and Li and lower in Si, Mn and Ti than those from any other pot. However, the three plants were established on the third seeding, and hence were less mature than the other plants. Their N content of 2.6% reflects this fact. N contents of other pots ranged from .72 to 1.36%, representative of Bromus in late seed production stages.

UPTAKE OF FERTILIZER¹⁵NO₃ AND¹⁵NH₄⁺

Enrichment in ^{15}N in plants adjacent to $^{15}NO_3^-$ and ¹⁵NH⁺-fertilized plots is reported in Table 8. Results are generally consistent with root zone estimates based on ¹⁵N uptake reported previously (Hunter et al. 1976b). Acamptopappus shockleyi and Ambrosia dumosa picked up no ¹⁵N outside the plot boundaries, indicating root spread of less than 2 m. Two Ceratoides lanata picked up some ¹⁵N from within one m of NH_4^+ -treated plots, but none from farther than 3 m. Two Ephedra, one 4 m and one 22 m from the plots showed enrichment. Four Larrea tridentata, all within 7 m of ¹⁵NO₃-treated plots, showed significant enrichment. Three Lycium andersonii, the farthest 7 m from a plot, and one Yucca schidigera at 15 m were enriched in ¹⁵N. These distances compare with the 1- to 6-m radii calculated from plants situated inside the plots.

The Ephedra nevadensis and Yucca schidigera showing uptake from 22 and 15 m, respectively indicate either unusual analytical error or distorted distribution of roots.

Plant samples from within the plot areas showed a concentration of ¹⁵N in new structures (Table 9). Nevertheless, old stems and dead wood showed significant dilution with ¹⁵N, indicating presence of a mobile N fraction of significant size. In all species but Ambrosia, live leaves showed the greatest enrichment, closely followed by fruit, flowers and new stem. Old live stem showed the least enrichment of live tissue, no doubt due to dilution with non-labile ¹⁴N. In Ambrosia, fruits rather than live leaves were most enriched.

Dead leaves were considerably enriched, reflecting their short life span. Essentially all dead leaves were produced after ¹⁵N fertilizer was applied. Enrichment of dead wood was of a low level, but generally higher than in 1975. It is not likely that such a high percentage of dead twigs was alive

Table 7. Enrichment of ¹⁵N in successive harvests of plants grown in Rock Valley soils which were amended with ground ¹⁵N-tagged tissues of desert plants

after treatment and it is possible that some enrichment occurs by nitrate movement from adjacent living tissues.

On nitrate-treated plots Bromus rubens showed less uptake than many shrub leaves. This may be due to growth of Bromus in shrub clumps higher in original NO₃ content, diluting the ¹⁵N. There may be significant dilution of Bromus nitrogen with seed N, as the plants were quite small. The fact that shrub leaves do show higher enrichment than Bromus suggests shrubs' labile N pools are not much larger than Bromus seed pools. This is in agreement with the Ambrosia data reported above, which showed little storage of N.

Enrichment was lower on plots treated with ¹⁵NH₄⁺ than on ¹⁵NO₃ plots, and lowest on the plot treated with N-Serve to inhibit nitrification. Acamptopappus, Ambrosia and Ceratoides were the most strongly enriched species, reflecting their smaller root zones. (Roots extending outside the 5-m² treated area cause reduced enrichment.)

Table 8. Percent ¹⁵N in tissues of plants near the four ¹⁵N-treated plots. Enrichment above background (0.27-0.43%) indicates distances over which nitrogen uptake can occur

Aca sho - Acamptopappue shockleyi, Amb dum - Ambrosia dumosa, Cer lan - Ceratoides lanata,
Eph nev - Ephsdra nevadensis, Eph fun - Ephsdra freera, Lar tri - Larrea tridentata, Lyc
and - Lydium andersonii, Yuc sch - Tucca s

As indicated in Table 10, levels of ¹⁵N in many plants or tissues have decreased, allowing a calculation of biological half-lives of N in desert shrubs. (Since half-lives below zero were not calculated, these values should not be used to approximate an average half-life.) The indicated values of 2-4 yr are shorter than we would have predicted, but as N is recycled through litter we expect an increase in half-life. Almost all plants on the plots treated with $^{15}NH_4^+$ have not reached a maximum enrichment, reflecting both later application and slower uptake due to soil adsorption and nitrification problems. Calculation of half-lives from the few samples which decreased in ¹⁵N enrichment is questionable, and we would expect significant changes at subsequent sampling times.

Soil analyses (Table 11) indicate that the majority of ¹⁵N present in the NO₃-fertilized plots remained in the NO₃ form 70 days later. The involved multiple soil assay and lack of precision in the newly instituted emission spectrographic ¹⁵N assay were particularly troublesome in these data. A direct spectrographic assay for soil ¹⁵N has been reported (Kanazawa and Yoneyama 1976) and is being investigated.

N BALANCE IN FIELD-INCUBATED SOILS

Changes in nitrogen fractions in field-incubated soils are recorded in Table 12. Some analyses have not yet been completed. Initial indications from these data are that soil nitrogen transformations in the Mohave Desert are quite slow and difficult to measure with our standard analyses.

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Table 9. Partition of ¹⁵N by tissue in Mohave Desert vegetation growing on plots treated Jan 29, 1975 (NO₃) and
Mar 21, 1975 (NH₄⁺) as of Jun 2, 1976. Background ¹⁵N (.27 to .43%) has not been subtracted

9a. ¹⁵NO₃-treated plots

9b. $15NH_4^+$ + ammonia treated

9c. $15NH_4^+$ + N-Serve-treated plot

| Plant | Live Leaves | Live Stem | New Stem | Flowers | Fruit | Dead Leaves | Deau Wood | Whole Shoot |
|---------------------|---|--------------|----------------|------------------------|--------------|----------------|-----------------------------------|----------------------|
| | | | | | | | | |
| | | | ÷. | | | | | |
| $\mathbf 1$ | 1.03 | 0.72 | 0.92 | | 0.88 | | 0.52 | |
| | 0.72 0.66 | 0.66 0.58 | 0.65 0.58 | \sim 0.60 | 0.83 | | 0.59 | $\ddot{}$ |
| | 0.74 | 0.72 0.67 | 0.96 0.74 | 0.97 \blacksquare | 1.16 0.74 | 0.74 | 0.65 0.53 | ù. |
| $\frac{1}{\lambda}$ | a | ż | | | ¥, | o e ×. | ٠ | c.t. |
| | 0.79 | 0.61 | 0.68 | | 0.83 | ú, | 0.43 | W) |
| | 0.64 | 0.59 | \overline{a} | | ÷. | ÷ | | ÷ |
| ÷ | \sim | 0.57 | 0.58 | | w | | ä, | ω. |
| $\mathbbm{1}$ | 0.57 | 0.41 | 0.49 | ÷ | × | ۰ | 0.42 | |
| | 0.46 0.45 | 0.43 0.35 | 0.46 0.47 | à. | ٠ | ż | 0.54 0.35 | ŝ |
| | $\bar{\gamma}^{\rm{SM}}_{\rm{NH}_4^+}$ + NSERVE $= 1$ Univ. \sim | 0.97 0.52 | 0.51 | 0.49 | 0.58 | 0.79 0.69 | $\frac{15}{9}$ N in Tissue _ ٠ | 0.50 0.44 |

* See footnote, Table 8; also, Bro rub = Bromus rubens, Gra spi = Grayia spinosa, Kra par = Kraneria parvifolia.

a"Shoot" for 1976 is the average of stem and live leaf values for 1977. It is a determination on leaf and stem ground together before analysis.

Table 11. Enrichment of ¹⁵N in soil fractions on ¹⁵NO₃-treated plots in Mercury Valley. Treatment on Plot 1 was 4 kg/ha NO₃-N, 33.6% enrichment, and on Plot 2, 1.76 kg/ha NO₃-N, 94.6% enrichment. Treated Jan 29, 1975 and sampled Apr 8, 1975

| Plot | Site | Depth | Kjeldahl digest to exclude NOT | | | Kjeldahl digest to inclue NO ₇ and NO ₂ | Soluble NO ₃ | Indicated $15N$ in $N03$ |
|----------------|----------------|-----------|-----------------------------------|-------------------|-------|--|----------------------------|-----------------------------|
| | | cm | 521 | gl5 _{Na} | 索 | 4^{15} N | ppm | 4^{15} N |
| ı. | | $0 - 15$ | 0.117 | 0.37 | 0.110 | 0.472 | 50 | 12 |
| | $\frac{1}{2}$ | $0 - 15$ | 0.055 | 0.45 | 0.056 | 0.580 | 8 | 61 |
| | | $0 - 15$ | 0.056 | 0.32 | 0.058 | 0.493 | | 343 |
| | | $15 - 30$ | 0.059 | 0.39 | 0.062 | 0.519 | | 78 |
| | | $30 - 45$ | 0.057 | 0.41 | 0.062 | 0.466 | 55555 | 64 |
| | 4 | 0.15 | 0.033 | 0.42 | 0.032 | 0.645 | | 61 |
| | | $25 - 30$ | 0.038 | 0.41 | 0.038 | 0.586 | | 131 |
| | | $30 - 45$ | 0.035 | 0.44 | 0.037 | 0.414 | | 35 |
| Controls | 1 | $0 - 15$ | 0.030 | 0.32 | 0.032 | 0.384 | | 17 |
| | | $15 - 30$ | 0.031 | 0.37 | 0.035 | 0.393 | | 43 |
| | \overline{c} | $0 - 15$ | 0.027 | 0.35 | 0.028 | 0.398 | mmm | 35 |
| | | $15 - 30$ | 0.031 | 0.38 | 0.034 | 0.412 | | 43 |
| $\overline{2}$ | $\frac{1}{2}$ | $0 - 15$ | 0.045 | 0.43 | 0.048 | 0.464 | 5 | 58 |
| | | $0 - 15$ | 0.105 | 0.48 | 0.097 | 0.557 | 60 | 13 |
| | \mathfrak{Z} | $0 - 15$ | 0.031 | 0.44 | 0.035 | 0.472 | | 26 |
| | | $15 - 30$ | 0.029 | 0.39 | 0.026 | 0.363 | S S | $11\,$ |
| | 4. | $0 - 15$ | 0.030 | 0.86 | 0.030 | lost | 10 | ۰ |
| | | $15 - 30$ | 0.025 | 0.53 | 0.024 | 0.710 | | 37 |
| | | $30 - 45$ | 0.019 | 0.50 | 0.020 | 0.527 | $\frac{8}{5}$ | 24 |
| Controls | 1 | $C - 15$ | 0.020 | 0.33 | 0.022 | 0.434 | 25 | |
| | | $25 - 30$ | 0.021 | 0.31 | 0.023 | 0.400 | 148 | $\frac{3}{1}$ |
| | 2 | $0 - 15$ | 0.022 | 0.36 | 0.022 | 0.398 | 8 | 23 |
| | | $15 - 30$ | 0.021 | 0.34 | 0.020 | 0.451 | 3 | 26 |

⁸Averages of 2 analyses run at adifferent time from other samples. These are somewhat higher than the values used to calculate $15n\%$ in the NO₃ fraction.

There was essentially no loss of ammonia in the ammonia volatilization experiment after six months (Table 12a, b and c). However, with 2 x normal water there was a decrease in total N and NO; during the six months from October through March. It appears that some NO; was generated in dry and 1 x normal **NH,+** -enriched soils during that same period, but it is difficult to explain any changes in the dry soils during that period. Skujins (1975) suggested N-Serve is lost by volatilization from dry soil. Its effectiveness in preventing nitrification would thus have been very transient.

Two of the three samples stored wet in lab and refrigerator and harvested April 7, 1977 were flooded. There was apparent denitrification in them. However, negligible change in total nitrogen indicates only nitrate was lost, not ammonia or organic nitrogen.

There was a gradual increase in exchangeable $NH₄⁺$ in all soils which initially contained \leq 5 ppm. Data are available for the first six months only. This increase was not related to water treatment. In cases where considerable ammonia was added there is little evidence for loss by volatilization from April through September (Tables 12b, h and k).

The denitrification experiment (Table 12j, k and 1) suggests there was considerable loss of **NO;** and total nitrogen from watered soils from October through March, but no change from April to October. This is in contrast to increases in nitrate found in shrub soils which appeared to be related to mineralization of organic nitrogen and occurred throughout the year.

It should be emphasized that these results are preliminary and subject to change upon reanalysis of some samples and a closer examination of the data.

ACETYLENE REDUCTION

Acetylene Reduction by Roots

Ethylene production by roots of five small *Yucca schidigera* harvested January 27, 1976 is reported in Table 13. There is a strong association of ethylene production with the presence of acetylene, suggesting nitrogen fixation associated with *Yucca* roots. In three of the five plants, ethylene produced from 72-141 hr exceeded our rule of thumb for maximum endogenous production (100 ng/gdw/hr; Hunter et al., 1976b). However, the lag in production seen from 0-71 hr suggests the ethylene produced is a result of incubation conditions rather than nitrogen fixation in the field. Lichen crust taken at the same time showed essentially no acetylene reduction.

The above sample developed a considerable mold population as incubation continued. We repeated the experiment as reported in Table 14. Before incubation a subsample was split into stele and epidermis and these portions incubated separately.

The form of basal roots differed between small and large *Yucca.* In the small *Yucca* the root stele is enlarged for several inches from the base, while large *Yucca* did not exhibit this

In this experiment, started March 10, there is no convincing evidence for nitrogen fixation. The lag in ethylene production is still present, but a production rate greater than 100 ng/gdw/hr was not found in any of the 20 whole root samples. Incubation for the final 15 days without acetylene did not reduce ethylene production in the whole roots.

Table 15 reports ethylene production by roots of nine Mohave Desert species. Each value is an average of five, one incubated without acetylene. There was great variability among roots, but acetylene appeared to have no effect on ethylene production rates. On a scale of 5 (highest), the ethylene production by roots without acetylene varied from 2.08 *(Ambrosia dumosa* and *Lycium andersonii)* to 3.56 *(Krameria parvifolia).* If all ethylene were produced by reduction of acetylene these roots would all have ranked 5.

Rates of ethylene production increased approximately tenfold between February and May. Since these were incubated in the field this is likely due to a combination of factors, including temperature and physiological state of the root. Some species differences are also apparent.

Only one sample exceeded 100 ng/gdw/hr *(Yucca,* April 27). Hence we feel these results cannot be construed to indicate fixation of nitrogen. They represent endogenous production of ethylene by shrub roots.

Soil and Litter Acetylene Reduction

Ethylene production by soil and litter samples taken February 3 is reported in Table 16. There is a tendency toward increasing ethylene production with time, except .in lichen crust and litter. However, production rates are well below values indicative of significant fixation, except for the later incubation period in the single algal sample from the glasshouse.

The small amount of ethylene present in most vials led us to suspect ethylene contamination in one of the reagents. Table 17 shows there is a small amount of ethylene produced by wet litter with and without acetylene. Dry soil and dry litter adsorbed a small amount of ethylene.

With use of the high concentrations of acetylene it is easy to get a trace of contamination from poorly flushed syringes. However, a careful check of the gas chromatograph charts shows a trace of acetylene in all but one serum bottle, even where nothing was added to the bottle, and where there was no chance of syringe contamination. Hence we believe the serum bottle stoppers produce a trace of acetylene. Larger traces were associated with litter and soils.

In several experiments a loss of acetylene from bottles containing dry soils, especially those from bare areas, was noticed (Table 18). Water (15%) was added to two such samples. Over the next half hour, acetylene peak heights increased 48 and 42 % , indicating dry soils adsorb acetylene and release it upon wetting.

Table 12. Nitrate, exchangeable ammonium and Kjeldahl nitrogen concentrations in field-incubated soils, and changes in these fractions during a one-yr incubation

12a. Soil nitrate concentrations during the ammonia volatilization experiment

12f. Total soil N during the litter decomposition experiment

12g. Soil nitrate during the nitrification experiment

| | | | | | | | | NITRATE-N, $\mu_{G/J}$ + sd | | |
|------------|---------------------------------|-------|-------|-----------|--------------|-----------|-----------|-----------------------------|----------------|-----------|
| | Treatment | | | \circ | | 1 w k | $1 \nmod$ | 3 _{mo} | 6 mo | 1 yr |
| | Water Fertilizer | Soil | Other | Apr 7 | | Apr 14 | May 7 | Jul ₇ | Oct 12 | Apr 7 |
| \circ | $+14H_2$ | Bare | | $30 +$ | $\mathbf{1}$ | $21 + 1$ | $36 + 3$ | $23 + 5$ | $31 + 3$ | $44 + 12$ |
| 1X | $+{\rm NH_{\rm 1}}$ | Shrub | | 9 ± 3 | | $11 + 2$ | $17 + 3$ | $7 + 0$ | $14 + 2$ | $12 + 3$ |
| | $+!{\rm TH}_{\rm i_4}$ | Bare | | $31 + 2$ | | $33 + 2$ | $40 + 1$ | $20 + 1$ | $33 + 1$ | $36 + 10$ |
| ${\rm SX}$ | $+$ $\rm{H}_{\rm{L}}$ | Shrub | | $6 + 7$ | | $11 + 2$ | $15 + 1$ | $8 + 1$ | $17 + 1$ | $26 + 4$ |
| | $+\mathrm{NH}_4$ | Bere | | $26 + 2$ | | $26 + 5$ | $38 + 1$ | $15 + 3$ | $35 + 1$ | $17 + 15$ |
| | $+3\mathcal{H}_{\underline{b}}$ | Shrub | | $9 + 2$ | | $11 + 2$ | $16 + 1$ | $7 + 1$ | $19 + 3$ | $14 + 5$ |

12b. Soil exchangeable ammonium concentrations during the ammonia volatilization experiment

| | | | | | | | AMMONIUM-N, μ g/g \pm sd | | |
|---------|--|-------|---------|--|--|---|-----------------------------------|--|--------|
| | Treatment | | | $\overline{0}$ | J MK | 1 mo | 3 mo | 6 _{mo} | 1 yr |
| Mater | Fertilizer | Soil | Other | Apr 7 | Apr 14 | May 7 | $\overline{7}$ Jul | Oct 12 | Apr 7 |
| \circ | $-17H_{1}$ | Bare | | $2.0 + 0.6$ | | $8.3 + 1.1$ | $1.9 +$ 0.2 | $6.0 + 0.1$ | |
| | $+17H_1$ | Bare | Sterile | | 111. τ $+$ | 42 $+$ τ | $\bf8$ | 64 $\frac{1}{2}$ τ | |
| 1x | $- W\mathrm{H}_{\mathrm{Li}}$ | Bare | $^{+}$ | $62 + 16$ $1.0 + 0.0$ $40 + 4$ $3.6 + 1.2$ | 1.2 + 38 + 3.5 + 0.3 | $\frac{1}{1}$ 0.1 $\frac{1}{1}$ 6.5 1.5 47 | $\overset{57}{_{2,1}}$ 0.1 | $\overline{+}$ 6.8 2.4 | |
| | $+\text{NH}_\text{L}$ | Bare | N-Serve | | 4 | | 5 | $\overline{+}$ 64 $\overline{7}$ | |
| 2X | $-1. H_1$ | Bare | | | 1.4 | 2.8 | 2.4 0.3 | $6.5 + 1.2$ | |
| | $+1\mathrm{HH}_1^2$ | Bare | | $79 + 14$ | 5 $35 +$ | $\frac{1}{2}$ 3 50 | $\overline{3}$ | $68 +$ 5 | |
| | $(rerun)$ | | | Jul 7 | Jul 14 | $Au - 7$ | Oct 12 | Feb 18 | Jul 21 |
| \circ | $-17H_0$ | Bare | | $\begin{array}{c} 2.7 \pm 0.1 \\ 57 \pm 20 \\ 4.0 \pm 1.0 \end{array}$ | 3.5 ± 0.3 | 0.8 4.4 | 3.2 1.7 \div | - | |
| | $+iH_0$ | Bare | Sterile | | $\frac{50}{3.4}$ $\mathbf 1$ $rac{+}{+}$ | $+1+1+1+1$ 7 53 | $+1+1+1$ 67 $\frac{3}{0.6}$ | u, | |
| 1X | $-17H_h$ | Bare | $+$ | | 0.2 | 1.5 | 2.4 | u, | |
| | $\begin{array}{c} +\overline{134} \\ -\overline{134} \\ +\overline{134} \\ +\overline{134} \\ \end{array}$ | Bare | | $70 + 3.8 + 1.5$ $\mathbf{1}$ | 57 4.2 Ŧ τ | 60 3.9 5 | $60\n2.8$ \overline{c} | ÷ | |
| SX | | Bare | N-Serve | 0.5 | $-22 + 72$ 0.8 | 1.1 | $.87 + 53 + 1$ 1.4 | è. | |
| | | Bare, | | $73 +$ 8 | \tilde{c} | 6 56 \pm | 8 | ÷ | |

12h. Soil exchangeable ammonium during the nitrification experiment

12c. Total soil N during the ammonia volatilization experiment

12i. Total soil N during the nitrification experiment

12j. Soil nitrate during the denitrification experiment

12d. Soil nitrate concentrations during the litter decomposition experiment

12k. Soil exchangeable ammonium during the denitrification experiment

12l. Total soil N during the denitrification experiment

Process Studies

12e. Soil exchangeable ammonium concentrations during the litter decomposition experiment

 V_{11}

Yu

Yu

Yu.

 V_{21}

Table 13. Acetylene-associated ethylene production by roots of Yucca schidigera harvested Jan 27, 1976. Roots were incubated in an argon-O₂-CO₂ atmosphere with and without acetylene

| | | | | | Table 14. Ethylene production rates of Yucca schidigera | |
|--|--|--|--|--|---|--|
| | | | | | roots incubated in an argon-O ₂ -CO ₂ -acetylene atmosphere | |
| starting Mar 10, 1977. For all values $N = 10$ | | | | | | |

 B_{00xing} the final incubation period for whole roots, half the samples were incubated without acctylene. There was no significant difference in ethylene production by these samples.

⁸Fach value represents five replicates, one without acetylene.

Table 16. Ethylene production under acetylene-argon-O2-CO₂ atmosphere by surface soils from beneath shrubs in Rock Valley. Samples were taken Feb 3, 1976

| Sammle Site | n | | Ethylene production |
|-------------------------------|---------------|-----------------------------|---------------------|
| | | $0-13$ days | 13-26 days |
| | | - ng ethylene/gdw/hr + se - | |
| Intershrub open areas | 33 5 10 | $0 + 0$ | $.002 + .002$ |
| Lichen crust, Mercury Valley | | $3.5 + 1.8$ | $.06 + .04$ |
| Benenth Lycium andersonii | | $.02 + .01$ | $.09 + .02$ |
| Beneath Larrea tridentata | 10 | $.02 + .01$ | $.07 + .01$ |
| Bencath Ambrosia dumosa | 10 | $.08 + .03$ | $.10 + .02$ |
| Beneath Ephedra nevadensis | 10 | $.03 + .01$ | $.07 + .02$ |
| Beneath Krameris parvifolia | 10 | $.02 + .01$ | $.05 + .01$ |
| Benesth Lycium pallidum | 10 | $.03 + .02$ | $.04 + .01$ |
| Bencath Gravia spinosa | 10 | $.03 + .01$ | $.06 + .02$ |
| Beneath Salazaria mexicana | 10 | $0 + 0$ | $.04 + .02$ |
| Litter samples, 2 per species | 17 | $21. + 128$ | $19. + 11.$ |
| Algal crust from glasshouse | | | 81. |

Table 17. Ethylene concentration in serum bottles as affected by our standard acetylene reduction assay procedures. Each treatment was replicated three times

⁸Rate for litter samples averaged 3.7 ± 2.0 for the first 5 days, hence 33.5 for the next 8 days.

aone unit = 2ppm in the 125-ml serum bottles.

bpeak width at middle of chart.

Table 18. Ethylene production by Mohave Desert soils following one week's incubation in a glasshouse. Soils were incubated (with water) from Jul 2 to Jul 9, 1977 and flushed with argon-O₂-CO₂ and acetylene (.2 ml, 10% acetylene). Desert soils from Mercury Valley, sagebrush interspaces soil from Shoshone Mountain

| Sai' Size | % Water | | | Acetylene Peaks ^a | | | Ethylene Units ⁰ |
|---------------------------------------|---|-------------------------|-----------------------|------------------------------|---------------------|----------------------|--|
| Rodent mound | $\begin{array}{c}\n0 \\ 15\n\end{array}$ | 2.8 5.7 5.7 | 3.5 5.3 5.7 | 3.3 5.5 5.8 | $F1$ O O | $\circ \circ$ | $\begin{array}{c} 0 \\ 0 \\ 1 \cdot 3 \end{array}$ |
| Desert pavement | $\begin{array}{c}\n0 \\ 5 \\ 15\n\end{array}$ | 3.5 5.4 5.5 | 4.0 5.3 5.7 | 4.3 5.5 5.5 | $\circ \circ \circ$ | $\circ \circ$ | $\circ \circ \circ$ |
| Wash soil with lichen crust | $\begin{array}{c}\n0 \\ 5 \\ 15\n\end{array}$ | 5.5 3.5 5.5 | 3.3 5.0 5.3 | 3.0 5.5 3.3 | \circ 0.6 | $\frac{0}{0}$ 0.9 | $\begin{matrix} 0 \\ 0.3 \end{matrix}$ 1.3 |
| Mountain soil, sagebrash community | $15\,$ | 5.3 | 5.3 | 5.7 | | | 6.5 6.2 12.8 |

akcetylene peak width at 50 units on gas chromatograph charts.

bone unit is approximately 2 ppm, 125-ml serum bottles.

Table 19. Soil nitrogen concentrations and ethylene production by Mohave Desert surface soils. Five soils of each type were incubated from Nov 23 to Dec 2 in a glasshouse under argon- $CO₂$ -O₂-acetylene atmosphere

one unit is approximately 2ppm in 125-ml serum bottle.

Table 18 also shows soils from Shoshone Mountain sagebrush interspaces produced much more ethylene (near 1) ng/gdw/hr) than do Mohave Desert shrub community interspaces. The Shoshone soil was much darker in color, indicating greater organic matter content.

Table 19 reports results from a similar experiment performed between November 23 and December 2. In this case lichen crust was much more active, while soil from sagebrush interspaces (from Mine Mountain) showed no activity. Soil nitrogen fractions are also reported for the same samples. Inorganic nitrogen fractions for the February 3 soil samples (Table 16) are reported in Table 20. In both cases ammonium was enriched, but quite variable, under lichen crust. In fact, the highest exchangeable ammonium was found in samples adhering to lichen crust. Total nitrogen under lichen crust was also higher than would be expected for the bare desert pavement areas where it occurs.

⁸Mercury Valley

Table 21. Ethylene production by Mohave Desert (Mercury Valley) soils over the course of a single day. Soils were sampled and injected between 0500 and 0600 Jun 30, 1976, and analyzed throughout that day and the morning of Jul 1. Units are ethylene peak integrator units, approximately equal to 2 ppm ethylene

| \boldsymbol{z} 1 1.3 1.8 | \overline{c} 2.6.1.0 | $1\,$ | \mathbf{S} | $\frac{1}{2}$ | \mathfrak{S} | 3 | $\frac{1}{2}$ | | $\overline{2}$ |
|--|---------------------------|-----------------------|----------------------------|-----------------------------|-------------------------|---|---------------|-------------------------|----------------|
| | | | | | | | | | |
| | 0.9 | | $0.0 \quad 0.0$ 0.0 0.0 | | 1.5 1.6 1.8 2.3 | | 1.3 | 0.0 | 0.0 0.0 |
| $\begin{array}{cc} 1.7 & 1.3 \\ & 1.0 \end{array}$ | 1.8 1.5 | | $0.0 \t 0.0$ | | 1.6 1.6 0.8 1.2 1.8 | | | 0.0 | 0.0 0.0 |
| | | | | | | | | | 0.0 0.0 |
| Average + se $1.5 + .1$ | | | | | | | | | |
| | | 1.0 1.3 $1.4 + .2$ | | $0.0 \t 0.0$ $0.0 + 0.0$ | | | $1.5 + 0.1$ | 1.0 1.4 1.4 1.3 | $0.0 + 0.1$ |

Table 21 shows results of an attempt to duplicate Skujins' (1976) data which showed diurnal fluctuation in Rock Valley acetylene reduction. The use of 10 % C_2H_2 in air, rather than in Ar -CO₂-O₂ was an attempt to duplicate his techniques, although a larger serum bottle (125-ml) was used. We were not able to detect any ethylene production by soils, although Table 21 shows a significant level of contaminating ethylene in the acetylene.

Microorganisms

In February 1976, rinse water from litter samples was used to inoculate N-free agar plates in an attempt to isolate N-fixing bacteria. The plates were left too long, and were overgrown by a fungus, tentatively identified as a Pullularia species. Several experiments were run on the Pullularia isolates and further isolates from litter (Tables 22 and 23). All organisms positive for N fixation in Table 22 were tested in agar containing streptomycin and tetracycline, in an attempt to eliminate bacterial fixation. It is apparent that tetracycline severely inhibited most acetylene-reducing isolates, while streptomycin did not.

Table 22. Ethylene production (acetylene reduction) associated with organisms growing on Mohave Desert plant litter mixed with N-free agar, and with "fungi" isolated from such litter

| | Treatment ^a | | April 9 | | April 16 | | April 19 | May 15 | |
|---------------------|--------------------------------|-----------|--|---|----------------|--|--------------------------|-------------|--|
| | Acetylone ucary | Acetylene | Ethylene | Acetylene | Ethylene | Acetylene | Ethylene | Acetylene | Etyhlene |
| Aca sho litter | | \circ | | \circ | T | | | | |
| Aca sho litter | $+$ $+$ | 8.0 | | 7.5 | 0.25 | | $\overline{}$ | 7.0 | 1.0 |
| Yue sch litter | | \circ | | | | | T | \circ | $\mathbb T$ |
| Yuc sch litter | $\ddot{}$ $+$ | 8.0 | $\begin{array}{c}\n0 \\ 0 \\ 0\n\end{array}$ | $\mathbb T$ | 56 | | | $\mathbf T$ | 1.0 |
| Fungus from Atr con | | \circ | 0.4 | $\overline{}$ | \blacksquare | T. | 0.2 | \circ | $\begin{array}{c} 1 \\ 0 \\ 1 \end{array}$ |
| | ÷. \pm . | 5.5 | $\mathbf T$ | | | 6.0 | 0.1 | 3.5 | |
| Fungus from Aca sho | | \circ | α | $\overline{0}$ | $\frac{1}{1}$ | \circ | Ω | \circ | |
| | $+$ $+$ | 8.0 | 0.5 | | | 7.0 | 276 | 3.5 | 293 |
| Fungus from Amb dum | | \circ | | $\ddot{\rm o}$ | \circ | T | \circ | | |
| | $\ddot{+}$ $+$ | 6.0 | \circ | 6.5 | | 3.5 | | 4.5 | \circ |
| Fungus from Cerlan | | \circ | | \circ | α | $\mathbf T$ | 2.8° | \circ | |
| | $\frac{1}{1}$ $\frac{1}{x}$ | 6.0 | 25.5 | 5.5 | 185 | $\mathbb T$ | 138 | ÷. | $\overline{\mathcal{C}}$ |
| "Red mold" from | | | | | | | | | |
| Yue sch | | \circ | \circ | 0 | \circ | | | | $\overline{\mathbb{C}}$ |
| | $\ddot{}$ | 7.5 | 1.8 | | \blacksquare | \blacksquare | | | |
| Pullularia sp. | | \circ | | $\begin{array}{c} 1 & 0 & 0 \\ 0 & 1 & 1 \end{array}$ | Ω | $\begin{array}{c} \circ \\ \circ \\ \circ \end{array}$ | Ω | 0 | \circ |
| | $\frac{1}{1}$ | 5.0 | | | 246 | | 223 | \circ | 30 |
| Control agar | | \circ | \circ | | | | \mathbb{O} | | |
| | | 7.8 | \circ | | | 5.5 | \circ | 6.0 | |

aunits for scetylene are width of acetylene peak (mm) at 50 units on gas
chromatograph chart, and for ethylene the integrator units on the same chart.
One unit of ethylene is approximately 2 ppm in the 125-ml serum bottles

Table 23. Sensitivity of acetylene reduction associated with fungi isolated from Mohave Desert plant litter to the antibiotics streptomycin and tetracycline

Conditions of the conditions of the conditional conditions of the flushing with
Conditions of the conditions of the conditions of the flushing with
-0.-00. One ethylene unit represents ~ 2 ppm in the 25-ml air volume.

*Not filtered - contained insects.

The lag phase in these cultures is a common finding in growth of microbial colonies. Disappearance of acetylene in the most active cultures confirms the reality of acetylene reduction, but acetylene in the most active cultures confirms the reality of acetylene reduction, but acetylene reduction in less active cultures is not differentiated from endogenous ethylene production. These studies were not continued. Attempts to transfer Pullularia to agar containing nitrogen failed, with a number of other fungal contaminates overgrowing the medium.

NITRATE AND AMMONIUM IN RAIN

Table 24 shows nitrate and ammonium concentrations in rainwater. Although some contamination from dust was unavoidable, the calculated nitrogen input of 2.3 kg/ha is a reasonable upper limit.

CALICHE NITRATE

The three dissolved "caliche" samples contained 306, 398 and 621 μ g NO₃-N/g dissolved rock.

DISCUSSION

None of the data reported above are inconsistent with the nitrogen cycle model presented in Figure 4 (Wallace et al., in press). Minor changes might be indicated by the short half-lives found with ¹⁵N, e.g. a greater potential mineralization rate or nitrification rate. Similarly the low indicated rate of atmospheric N_2 fixation might be further decreased. Indications of the nitrogen balance experiment reported above are that denitrification and ammonia volatilization losses are small.

The Mohave Desert appears to have a total reserve of N on the order of 100 x annual plant uptake, with immediately available N sufficient for several years' uptake. Input in rain can supply about 10% of annual needs, with fixation of N_2 supplying perhaps 1% of the needs. Although a gradual net loss of N from the desert cannot be ruled out, an equilibrium state or net gain (as $NO₃$) seems most likely. The Chilean nitrate deposits and data of Boyce et al. (1976) suggest deserts may accumulate NO₃.

It appears from *Ambrosia* harvest data and enrichment of new tissues by 15NO; uptake that N for new growth is supplied by soil reserves, rather than plant reserves. This simplifies consideration of annual needs, which we have estimated from annual production on the basis that new growth depends on soil uptake of nitrogen.

The data indicating root zones with radii of up to 7 m for *Larrea* are consistent with findings that plants several meters distant from irrigated plots tended to be affected by irrigation (Hunter et al. 1976a). The alternative explanation was that horizontal water flow occurred over those distances in the vapor phase. Vollmer et al. (1975) have excavated *Larrea tridentata* roots for distances up to 1.1 m, and one *Krameria parvifolia* root was followed to 3 m. A photograph of a *Larrea* plant with a root several meters long is shown by Wallace and Romney (1972;123). Determining these root lengths by excavation is a formidable task in the very rocky soils of the northern Mohave Desert, and the ¹⁵N data, if expanded, would be quite useful.

The short half-lives of 3-4 yr indicated those both in pots in the glasshouse and in the field need clarification. The greenhouse pots received much more water, at average temperatures warmer in winter and cooler in summer than in the field. Under such conditions denitrification, volatilization and plant uptake are unrelated to the field situation. Nevertheless, the non-harvest loss of 15N appears evidence for loss either by denitrification or ammonia volatilization, or immobilization as organic **N.** In the field the half-life is essentially a half-life for availability, not residence time. *As* the 15N is bound up in plant tissues, lost as litter or equilibrated with other soil fractions, it is temporarily lost from the available N pool, and these transformations reduce the apparent half-life as measured by falling plant tissue concentrations. We feel after equilibration in the system the apparent half-life will show a considerable increase.

Process Studies

Our acetylene reduction assays show very consistently low or zero fixation of atmospheric ammonia. Almost all ethylene produced, except by microorganisms cultured on N-free agar, can be explained by reactions other than acetylene reduction or by contamination of the acetylene supply. For the few root samples in which a positive acetylene reduction result was found, the lag before significant ethylene production suggests the rates measured are not representative of field conditions. Indeed, given the long incubation periods, the results may be due to attack by decomposing organisms rather than normal rhizosphere microbes.

Nitrogen in rain is the largest apparent input of **N** to the Mohave Desert. However, there is some evidence that the source of this **N** is the desert itself. Yaalon (1964) found **NH,+** concentration in rain increased with soil temperature over calcareous soils, and suggested volatilization was the source. In a study of ammonia in rain in the United States, Junge (1958) attributed increases over the western states to calcareous soil.

Woodmansee (1976) considers rain and dust and direct absorption of **NH,** by plants to be the principal sources of nitrogen input to the short grass prairie.

Table 12 suggests **NH,** volatilization is insignificant. This is similar to results of Skujins (1975, 1976) who found **NH,** volatilization losses of less than 1% of soil NH⁺ in three weeks, both with and without NH₄⁺ fertilizer.

There are several indications that NO; builds up to high levels in Mohave Desert soils. Results presented by Romney et al. (1973), Hunter et al. (19756) and the caliche studies above have found concentrations from 100 to 1000 μ g/NO₃-N/g. In these cases NO₃ behaves as any other salt. Such concentrations would be unlikely if denitrification occurred to a significant degree, or if NO; were leached below the

Figure 4. Nitrogen cycle model-northern Mohave Desert. Units are kg N/ha for compartments (boxes) and kg N/ha-yr for fluxes (arrows).

depth sampled. The association with caliche and the presence of high concentration levels of NO; between 30-90 cm suggest to us a concentration at the lower drying front of Mohave Desert soils.

It is possible that over geological time periods nitrate may replace CO_3 ^{$=$} in the carbonate-silicate "caliche" deposits, leading to nitrate deposits or high nitrate levels in ground water.

Concentration of NO; and other salts under shrubs may be related both to water movement toward the drier soils under shrubs and collection of litter and its associated salts (Romney et al., in press). Hanawalt and Whittaker (1977) found mass flow of ions to roots greater than uptake for most salts in the Mohave Desert. If that situation holds for NO; as well, the concentration under shrubs and depletion of nitrogen in bare areas might be partially explained by mass flow of water toward roots. Validation site studies (Turner et al. 1975; Maza and Turner 1976) consistently show a gradient of water potential between open and shrub areas which would cause flow of water (and salts) toward the shrubs.

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