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Model for Predicting Simultaneous Distribution of Salt and Water in Soils

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MODEL FOR PREDICTING SIMULTANEOUS

DISTRIBUTION OF SALT AND

WATER IN SOILS

by

Satish C. Gupta

A dissertation submitted in partial fulfillment of the requirements for the degree

of

DOCTOR OF PHILOSOPHY

in

Soil Science and Biometeorology

(Soil Physics)

Approved:

UTAH STATE UNIVERSITY Logan, Utah

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Satish C. Epyta

Satish C. Gupta

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ABSTRACT

Model for Predicting Simultaneous Distribution

of Salt and Water in Soils

by

Satish C. Gupta, Doctor of Philosophy Utah State University, 1972

Major Professor: Dr. R. J. Hanks
Department: Soil Science and Bio Soil Science and Biometeorology

Knowledge of water and salt movement in soils is necessary for development of a management scheme for controlling the quality of irrigation return flow. A computer model was developed to predict the distribution of water and salts in the root zone under varying initial and boundary conditions. The model consists of water flow and salt flow sub-models. The water flow sub-model considers the numerical approximation of the general water flow equation with modification for water loss by evapotranspiration (and thus root extraction). The salt flow sub-model considers the mass flow of salts, chemical exchange, precipitation or dissolution of $CaCO_{2}$, $CaSO_{4}$, and formation of undissociated Ca and Mg sulphate.

The model was tested under laboratory and field conditions by comparing predicted values with experimental measurements. Satisfactory agreement was noted for the water content distribution in almost all the experiments. The model yielded approximately correct values of total salt distribution in the field and one of the column experiments. The agreement between the measured and predicted values for the two

other column experi ment s was poor . The poor agre ement seems to result from the irregular dissolution of the applied powdered salts. The distribution of individual ions was not accurately predicted by the model. The disagreement between the predicted and measured values was large at high salt concentration. Complex ion formation, insufficient description of exchange and activity coefficients at high salt concentration are suggested for this lack of agreement. Further development and field testing of the model are needed.

(112 pages)

INTRODUCTION

Public awareness of environment degradation has created an urgent need for re-evaluation of management techniques in various industries including agriculture. It has been claimed that return flow from various irrigation projects is one of the major contributors to the quality deterioration of streams. Government agencies have given high priorities to the problem of water quality of irrigation return flow and possible means for its control. A project was initiated on the Hullinger Farm near Vernal, Utah in 1970, to develop and field test a scheme to predict and control the quality of irrigation return flow. In this scheme it is necessary that the process of simultaneous transfer of salts and water in soils be understood. The dissertation of this writer involved this part of the project.

The transport of chemical substances through a porous medium in either liquid or vapor depends upon the combined action of diffusion and mass flow. However, chemical interactions such as adsorption, fixation, precipitation, and breakdown or decay adds to the complexity in transport processes. One of the simple systems which is not affected by the above chemical effects is the transport of non-adsorbed salts like chloride and nitrate. Bresler and Hanks (1969) have successfully built a computer model to describe the movement of these salts in the soil profile, under varying boundary conditions. Since a great number of solutes react with the soil it is necessary to deal with this problem to represent more closely the real situation.

Fortunately, in recent years, with the wide spread use of digital computers and better understanding of chemical processes in soils, there are methods available which show promise of being capable of handling the flow of interacting ions in soils.

The purpose of this study was to develop and field test a model for predicting the concentration of adsorbed ionic species in soils under unsaturated flow with root extraction. The following specific **chemical processes are considered.**

1. Precipitation or dissolution of gypsum, calcite.

2. Formation of undissociated Ca and Mg sulphate.

3. Interaction between ions in solution and solid phase. The ionic species considered are Ca⁺⁺, M_g⁺⁺, Na⁺, Cl⁻, HCO₃, and $\mathrm{SO}_{\Lambda}^{\equiv}$.

LITERATURE REVIEW

Column Chromatography

Most of the models developed for tracing salt disturibution in soils are based on the laws of conservation of mass. They state that the amount of salt added by water applied to the soil layers, minus the amount leached out and the amount absorbed by plants is equal to the net increment (positive or negative) of salts in the soil layer.

Any attempt to gather information on the vertical transport of different ions or salt solutions through the soil results in a mass of chromatographic theories. Two different approaches can be defined in the literature. The first one is based on the kinetic process called the "rate theory" (DeVault, 1943; Hiester and Vermeulen , 1952; and Lapidus and Amundson, 1952). The second one is the plate theory of Glueckauf (1949), Thornthwaite, Mather, and Nakamusa (1960), Dutt et al. (1971, and Bresler (1967) in which the height of a plate in the column is the unit of calculation. Historical development of the two different schools of thought will be reviewed separately in the following sections.

Rate theory

One of the simplest rate theories is that of DeVault (1943) . It is also described as the equilibrium chromatography. It requires that the penetrating solution move through the porous medium at such a rate that a dynamic equilibrium between the ions in solution and adsorbed phase shall be maintained. The theory starts with a material balance

over a cross sectional layer of the column of thickness dz:

$$
\frac{\partial c}{\partial z} + \alpha \frac{\partial c}{\partial v} + \frac{\partial E}{\partial v} = 0
$$
 [1]

where c is the concentration of solute in the solution phase. E is the concentration of solute in the solid phase, z is the distance from the top of the column, α is the pore or void fraction of the column and v is the volume of the solution fed to the column. Under saturated flow the general solution of this equation is

$$
z = g(c) + \frac{v}{\alpha + \phi f'(c)}
$$
 [2]

where ϕ is the amount of adsorber per unit of length, $f'(c)$ is the derivative of $f(c)$ with respect to c, $f(c)$ is the adsorption isotherm defined in such a way that $E = \phi f(c)$ and $g(c)$ is any function determined by the initial distribution of solute through the column.

Rible and Davis (1955) applied this theory with some success to **predict ion distribution in soils. The theory is less involved** mathematically but is limited in application to soil because of the assumption of instantaneous equilibrium and negligible channeling.

Hiester and Vermeulen (1952) started with another material balance equation

$$
-\left(\frac{\partial c}{\partial u}\right)_v = \left(\frac{\partial E}{\partial v}\right)_u + \alpha \left(\frac{\partial c}{\partial v}\right)_u
$$
 [3]

where u is the bulk packed volume of the column $(u\alpha = void volume)$ up to point z. Their work was the extension of work started by Thomas (1944) who took account of the rate of exchange by second order kinetics. The starting point is

$$
A^+ + BX \rightarrow B^+ + AX
$$
 [4]

and the rate equation being

$$
\frac{\partial E}{\partial t} = k \left[c(E_T - E) - \frac{E}{K} (c_0 - c) \right]
$$
 [5]

where A^+ , B^+ are cations and X is the exchanger, E_{T} is cation exchange **capacity, c 0 is the total cation concentration in solution, k is the** specific rate factor, and K is the equilibrium constant. They further **defined dimensionless parameters, solution capacity parameter T, column capacity parameter s, and equilibrium parameter r,** *to* **reduce** equation [3] to

$$
-\left[\frac{\partial(E/E_{\text{T}})}{\partial \text{T}}\right]_{\text{S}} = \left[\frac{\partial(c/c_{\text{o}})}{\partial \text{s}}\right]_{\text{T}}
$$
 [6]

and equation [5]

$$
\left[\frac{\partial (E/E_T)}{\partial T}\right]_S = \left[\frac{c}{c_o} \left(1 - \frac{E}{E_T}\right) - r \frac{E}{E_T} \left(1 - \frac{c}{c_o}\right)\right]
$$
 [7]

Hiester and Vermeulen (1952) have provided the graphical and numerical solution to the equation $[6]$ and $[7]$ in terms of parameter T , s, and r.

Bower, Gardner, and Goertzen (1957) tested this theory in the soil system. They found a reasonably good agreement between the theoretical and experimental values for the distribution of dissolved and exchangeable ions in the soil solumn as a function of depth and volume of solution applied. Gardner and Brooks (1957) distinguished between immobile and mobile salt moving with the same velocity as the leaching front. They adopted and tested the theory of Hiester and Vermeulen (1952) in laboratory column and field plots of Pachappa sandy loam. Agreement between the predicted and experimental values was found to be satisfactory.

The model of Hiester and Vermeulen differs from the proceeding model of DeVault in that rate dependent processes are considered in lieu of the assumption of equilibrium. However, both the models ignore the dispersion of salts.

The third model that is based on kinetics is by Lapidus and Amundson (1952). They have developed a model which takes into account the dispersion in addition to the mass flow. Previous work of Nielsen and Biggar (1962) has shown this model as the most satisfactory of all models investigated for predicting the spreading of a non-interacting **solut e, in porous media, where spreading results from diffusion and** dispersion. When exchange is also considered the material balance **over a layer dz is,**

$$
D\frac{\partial^2 c}{\partial z^2} = \overline{v} \frac{\partial c}{\partial z} + \frac{\partial c}{\partial t} + \frac{1}{\alpha} \frac{\partial E}{\partial t}
$$
 [8]

where D is the dispersion coefficient. Depending upon the boundary conditions the equation can be solved analytically (Nielsen and Biggar, 1962) or numerically (Lai, 1970). Some of the assumptions implied in the above model are that the velocity profile can be represented by an average \overline{v} , the diffusion coefficient is constant, equilibrium between the two phases is established and there exists some relationship between the ions in solution and the exchanger. A comparative study of three models, DeVault (1943), Hiester and Vermeulen (1952), and Lapidus and Amundson (1952) was reported by Biggar and Nielsen (1963) using Oakley sand. They concluded that all the theories were generally inadequate to describe the experimental values. The lack of agreement was attributed to the inadequate description of exchange, the

use of the average value of the flow velocity and the diffusion coefficient .

Plate theory

In the plate theories the column is regarded as being divided into a large number of segments or plates. Within each plate the concentration is considered to be uniform both in sorbent and liquid phases, the two concentrations being assumed to be at equilibrium. It is immaterial whether an exchange process, diffusion process, or any other process is envisaged as the main dispersion process. It is also implied that there is only one unit of length to which this definition applies. If the height of the plate is too long, the concentration may not be regarded as uniform and if it is too short no equilibrium is possible between **the two concentrations.**

One of the plate theories that has been extensively tested in the soil system is that of Glueckauf (1949). The starting equation is,

$$
\left(\frac{\partial f(c)}{\partial v}\right)_z + \left(\frac{\partial c}{\partial z}\right)_v - \frac{\Delta}{2} \left(\frac{\partial^2 c}{\partial z^2}\right)_v = 0
$$
 [9]

where \triangle is the plate height. In 1956, van der Molen applied this theory to the desalinization of soils under the influence of Dutch climate (mean annual precipitation about 70 em). The solution of equation [9] for a homogeneous saline profile at large values of N and in the case of a linear adsorption isotherm, may be represented by

$$
c = \frac{c_o}{2} \operatorname{erfc} \frac{p-1}{\sqrt{2 p}} \sqrt{N} \tag{10}
$$

where N is the number of depths above a distance z, $p = \frac{v}{\alpha \rho d}$,

erfc n = 1 - $\frac{2}{\pi}$ $\int_0^\pi e^{-u^2} du$. ρ is the density of the soil, and d is the \circ **distance from the soil surface. Some general agreement was found in** theoretical and observed values. Dyer (1965) studied the distribution of chloride and nitrate ions in adjacent irrigated and non-irrigated areas and observed a close fit of theoretical values with the observed **ones.**

Finally, there are two other very practical theories that have **nevertheless retained the characteristics of chromatographic transport.** Both the theories consider the fixed plate height. Thornthwaite et al. (1960) and Frissel and Poelstra (1964) have described the transport of **strontium through soils. Their method is based on Hartin and Synge** (1941) theory except that the plate height is fixed and it is assumed that 0.1 part of Sr in each layer is leached downward to the next layer, for every unit of the leaching solution added to the soil. If at the beginning of leaching only one layer is loaded, the concentration in the nth layer c_n is found from

$$
c_n = \frac{t!}{(t'-n+1)!(n-1)!} 0.1^{(n-1)} 0.9^{(t'-n+1)} c_0
$$
 [11]

where c_{o} is the total concentration in the first layer and t' is the number of leaching cycles.

The other approach which has received wide spread attention is by Dutt (1963) and his co-workers. He has used his method for calculating the quality of water percolating through soil containing gypsum. The concentration of salts at any depth and time is given by

$$
c_{i}^{j} = c_{i}^{j} - 1 \left(\frac{\theta s_{i} - \theta_{i}^{o}}{\theta s_{i}} \right) + c_{i}^{j} - 1 \left(\frac{\theta_{i}^{o}}{\theta s_{i}} \right)
$$
 [12]

where i and j are depth and time, respectively. θ is the moisture content and θS_i is pore volume at any depth. c_i^j is then corrected for solubility of minerals and exchange with the soil.

The advantage of this type of approach is that it is possible to **introduce such factors as solubility of minerals, etc. Since exchange constants are used it means no linear adsorption isotherm is necessary.** The main drawbacks of Dutt (1963) model are that the process is **dlscontinuous and unknmvn dispersion is present even when physical** dispersion is ignored in the model.

Review of the different models applied to the soil for describing the movement of salt are discussed by Frissel et al. (1967) and Biggar and Nielsen (1963). For the most part previous investigators have used constant flow velocities. A notable exception is the work of Bresler and Hanks (1969) who describe the numerical technique for simultaneous flow of water and salt in unsaturated soils and allow for time dependent velocities. This work was essentially a combination of Bresler (1967) model for salt flow, and Hanks and Bower (1962) model for water flow. The model starts with a material balance equation

$$
\left[\frac{\partial(\theta c)}{\partial t}\right]_z = \left[\frac{\partial}{\partial z}(\theta D \frac{\partial c}{\partial z}) - \frac{\partial(\overline{v}c)}{\partial z} + S\right]_z
$$
 [13]

where \bar{v} is the volumetric flux of water given by Darcy's law, t is the time, S is the sink or source term due to the solubility of mineral or exchange between solid and solution phase. Equation [12] is similar to

the equation given by Lapidus and Amundson (1952) . The present model of Bresler and Hanks (1969) contains the important features of rate as well as plate theory. The plate height is variable with depth but is constant with time. The model ignored the dispersion and sink or source **term. However, a critical examination of the numerical method indicates** a tendency for the concentration profile to spread for non-interacting salts rather than have a sharp profile, thus indicating a "built in" dispersion in numerical approximation like Dutt (1963). The model has been tested in the laboratory and gave values which agreed well with the experimental results. The theoretical development of **this model and its modifications are described in the next section.**

THEORETICAL CONSIDERATIONS

The present model is essentially the combination of the model for water and salt flow by Bresler and Hanks (1969) and Dutt et al. (1971) models for solubility of minerals and exchange between the solution and solid phase. The essential features of both models are discussed in the following paragraphs. One dimensional flow is considered in the model.

Water Flow Model

The basic water flow equation for one dimension is taken as

$$
\frac{\partial \theta}{\partial t} = -\frac{\partial \overline{v}}{\partial z} + A(z)
$$
 [14]

where θ is the water content and $A(z)$ is the plant root extraction function. The theoretical development of $A(z)$ have been discussed by Nimah (1972) and will not be discussed in the present derivation $(A(z) = 0)$. The volumetric flux of water \overline{v} is given by Darcy's law

$$
\bar{\mathbf{v}} = - \mathbf{K}(\theta) \begin{bmatrix} \frac{\partial \mathbf{H}}{\partial \mathbf{z}} \end{bmatrix}
$$
 [15]

where $K(\theta)$ is unsaturated hydraulic conductivity and H is the hydraulic head defined as

$$
H = h + z \tag{16}
$$

where h is the pressure head. A numerical solution of equation [14] is given by Hanks and Bower (1962) and a review of solutions by

Freeze (1969). The numeric form of equation [14] is

$$
\frac{\theta_1^j - \theta_1^{j-1}}{\Delta t} = \frac{\begin{bmatrix} h_1^j - 1 + h_1^j - 1 + 2G - h_1^{j-1} - h_1^j \end{bmatrix} k_1^j - 1/2}{2(\Delta z_1)^2} - \frac{\begin{bmatrix} h_1^j - 1 + h_1^j + 2G - h_1^j - 1 - h_1^j + 1 \end{bmatrix} k_1^j - 1/2}{2(\Delta z_1)^2} \tag{17}
$$

where G is the gravitational term. $G = \Delta z_i$ for vertical infiltration down, $G = -\Delta z_i$ for vertical infiltration up and $G = 0$ for horizonal infiltration. Δz_i is variable in the present model. The derivation of equation [17] assumes a unique relationship between pressure head, h, and water content, θ . If the assumption holds then it is possible to write a relationship between θ and h,

$$
\frac{\theta_1^j - \theta_1^j - 1}{\Delta t} \approx \frac{\left[h_1^j - h_1^j - 1 \right]}{\Delta t} \ B_1^j - 1/2 \tag{18}
$$

where B is the specific moisture capacity defined as

$$
B_{i}^{j} = 1/2 = \left(\frac{\partial \theta}{\partial h}\right)_{i}^{j} = 1/2
$$
 [19]

Substitution of equation [18] reduces equation [17] to

$$
\frac{h_1^j - h_1^{j-1}}{t} = \frac{\left[h_1^j - 1 + h_1^j - 1 + 2G - h_1^{j-1} - h_1^j \right] \kappa_1^j - 1/2}{2(\Delta z_1)^2 h_1^j - 1/2}
$$

$$
- \frac{\left[h_1^j - 1 + h_1^j + 2G - h_1^j - 1 - h_1^j + 1 \right] \kappa_1^j - 1/2}{2(\Delta z_1)^2 h_1^j - 1/2}
$$
 [20]

An equation similar to equation [20] can be written for each depth increment involving unkown of $h^j_{i}(i = 1, 2, 3, \cdots, n)$. The boundary condition supplies the value of $h^{\frac{1}{j}}_0$ and $h^{\frac{1}{n}}_n + 1$. Equation [20] can also be expressed as

$$
-AA_1 h_1^j - 1 + BB_1 h_1^j - CC h_1^j + 1 = DD_1
$$
 [21]

where

$$
AA_{1} = \frac{\Delta t}{2(\Delta z_{1})^{2}B_{1}^{j}} - \frac{1/2}{1/2}
$$
\n
$$
BB_{1} = 1 + \frac{\kappa_{1}^{j} - 1/2}{2(\Delta z_{1})^{2}B_{1}^{j}} - \frac{1/2}{1/2} + \frac{\kappa_{1}^{j} - 1/2}{2(\Delta z_{1})B_{1}^{j}} - \frac{1/2}{1/2}
$$
\n
$$
CC_{1} = \frac{\Delta t}{2(\Delta z_{1})^{2}B_{1}^{j}} - \frac{1/2}{1/2}
$$
\n
$$
DD_{1} = h_{1}^{j} - 1 + \frac{\kappa_{1}^{j} - 1/2}{2(\Delta z_{1})^{2}B_{1}^{j}} - \frac{1/2}{1/2} + \frac{\kappa_{1}^{j} - 1/2}{2(\Delta z_{1})^{2}B_{1}^{j}} - \frac{1/2}{1/2} + \frac{\kappa_{1}^{j} - 1/2}{2(\Delta z_{1})^{2}B_{1}^{j}} - \frac{1/2}{1/2}h_{1}^{j} - \frac{1}{1} + 2G - h_{1}^{j} - 1\right)
$$
\n
$$
- \frac{\kappa_{1}^{j} - 1/2}{2(\Delta z_{1})^{2}B_{1}^{j}} - \frac{1/2}{1/2} \left[h_{1}^{j} - 1 + 2G - h_{1}^{j} - 1 \right]
$$

$$
\hbox{when boundary conditions are substituted, n equations result in}\\
$$

tridiagonal matrix

~B1 - ee1 0 hl DD1 ec2 h2 DD2 l-AA, "> - -~3.B~3.-cc3 h3 DD3 -AA BB h lnn'n n n n [22]

Equation [22] is solved for $h^{\hat{j}}$ by the regular techniques of s olving tridiagonal matrix (Carnahan, Luther, and Wilkes, 1969). The water content, θ , is estimated from the relationship between θ and h.

Salt Flow Model

The rate of flow of salts at any plane in the direction of flow may be given by the equation [12]

$$
\left[\frac{\partial(\theta c)}{\partial t}\right]_z = \left[\frac{\partial}{\partial z} (\theta D \frac{\partial c}{\partial z}) - \frac{\partial(\overline{v}c)}{\partial z} + S\right]_z.
$$

The first term on the right, in the above equation, represents the contribution from diffusion to the flow of solute and the second term represents the contribution from viscous flow. S is the sink or source term due to solubility of minerals and exchange of ions in solution with solid phase. Each component of equation [12] is discussed separately.

Mass flow of salts

If the dispersion is absent and no sink or source exists, the flow of salt is due to the mass flow of water expressed as

$$
\left[\frac{\partial(\theta c)}{\partial t}\right]_z = -\left[\frac{\partial(\overline{v}c)}{\partial z}\right]_z \quad . \tag{23}
$$

Numerical approximation of equation [23] leads to

$$
\left[\begin{array}{cc} \theta_1^j c_1^j - c_1^j \end{array}\begin{array}{c} -1 \\ \theta_1^j \end{array}\begin{array}{c} -1 \\ \theta_1^j \end{array}\right] = -\frac{\left[\begin{array}{cc} -1/2 \\ \nabla_1^j \end{array}\begin{array}{c} -1
$$

with the approximation, c_1^j $1/2$ = c_1^j $-$ 1 and c_1^j $1/2$ = c_1^j 1 , equation [24] reduces to

$$
c_1^j = \left((\overline{v}_1^j - 1/2 c_1^j - 1 - \overline{v}_1^j - 1/2 c_1^j - 1) \frac{\Delta t}{\Delta z} + c_1^j - 10 \beta_1^j - 1 \right) / 0 \frac{1}{i} . \quad [25]
$$

Equation [25] and its modification are used to compute the mass flow of salts due to water.

Dispersion of salts

If there is no dispersion, there should be piston flow of salts and sharp boundary in the salt distribution should exist at the wetting front. Since the numerical approximation involves the mixing of solutions and then averaging over a new water content (equation [25]) a diffuse salt boundary exists at the wetting front. Although in the present model dispersion is ignored explictly, the mixing of salt indicates a "built in" dispersion in the numerical method of salt flow.

Sink or source term

The concentration of salts at each depth is modified due to the chemical reactions like precipitation or dissolution of minerals and exchange between ions in solution and soil matrix. Both these processes contribute to the source or sink term in equation [12].

Dissolution or precipitation of gypsum. A slightly soluble salt often present or added to the soil is gypsum. An equation relating gypsum to other constituents in soil is

$$
\text{Cas}_4. \quad 2\text{H}_2\text{O} \stackrel{\rightarrow}{\leftarrow} \text{Ca}^{++} + \text{SO}_4^- + 2\text{H}_2\text{O} \quad . \tag{26}
$$

The solubility of gypsum is described by the solubility product constant concept

$$
Ksp = {}^{a}Ca {}^{a}SO_{4} = {}^{c}Ca {}^{c}SO_{4} \ Y^{2} = 2.4 \times 10^{-5}
$$
 [27]

where Ksp is the solubility product constant, a is the activity of the ions designated, γ is the activity coefficients of divalent ions ${\gamma_{\mathrm{SO}_4}}$), and c is the equilibrium concentration of ions designated

which are defined further as follows.

Let x moles per liter of Ca^{++} and $SO_{\Lambda}^{=}$ that dissolve or precipitate and c_{Ca}^{0} , $c_{SO_{4}}^{0}$ are the initial molar concentration of ca^{++} and SO_{4}^{+} , respectively. Then the change in relative composition of Ca^{++} and SO_{Λ}^{\equiv} is

$$
c_{Ca} = c_{Ca}^{O} + x \qquad [28]
$$

$$
c_{S0_{4}} = c_{S0_{4}}^{0} + x . \qquad [29]
$$

Combining equation [28] and [29] with equation [27], results in equation of the form

$$
x^2 + Bx + C = 0
$$
 [30]

where

$$
B = {c_{Ca}^{o} + c_{SO_{4}}^{o} \over c_{Ca} + {c_{SO_{4}}^{o} \over c_{SO_{A}} - Ksp/\gamma^{2}}.
$$

Equation [30] can be solved for x.

Undissociated Ca and Mg sulphate. In addition to the dissolution or precipitation of gypsum, the CaSO₄ - Ca⁺⁺, SO₄ - H₂O system involves the formation of undissociated $caso_{\lambda}$. The dissociation constant K [CaSO^o] of ion-pair is defined as

$$
K_{\text{[Cas}\Omega_4^{\circ}]}} = \frac{c_{\text{Ca}} \cdot c_{\text{SO}_4^{\circ}\gamma^2}}{c_{\text{Cas}\Omega_4^{\circ}\gamma}}
$$

where c CaSO $^{o}_{\Lambda}$ is the molar concentration of the ion-pairs and γ for ionpairs is taken as unity.

Let x_1 be the moles per liter of ca^{++} and SO^+_{Λ} which forms undissociated CaSO₄. If the initial concentration of CaSO₄ ion-pair is $c_{\texttt{CASO}_{\tilde{\Lambda}}^O}^{\texttt{O}}$ then the change in concentration will be

$$
c_{Ca} = c_{Ca} - x_1 \tag{32}
$$

$$
{}^{c}SO_{4} = {}^{c}SO_{4} - x_{1}
$$
 [33]

$$
{}^{c}c_{a}so_{4}^{o} = {}^{c}c_{a}so_{4}^{o} + x_{1}
$$
 [34]

when equation [32), [33), and [34) are combined with equation [31), rearrangement yields an equation of the form

$$
Ax_1^2 + Bx_1 + C = 0.
$$
 [35]

The chemistry of undissociated $MgSO_4$ is similar to $CasO_4$ and results in an equation similar to equation [35), where

$$
A = \gamma^{2}
$$

B = - $\left[K_{[CasO_{4}^{o}] \text{ or } [MgSO_{4}^{o}] + \gamma^{2} \text{ c}_{Ca \text{ or } Mg + \gamma^{2} \text{ c}_{SO_{4}}^{o}}\right]$

$$
c = \gamma^2 \, c_{\text{Ca or Mg}}^o \, c_{\text{SO}_4}^o - \kappa_{[\text{CasO}_4^o]}^o \text{ or } [\text{MgSO}_4^o] \, c_{\text{CasO}_4^o}^o \text{ or } \text{MgSO}_4^o \,.
$$

Equation [35] can be solved for x_1 . When the system contains gypsum, the undissociated CaSO, becomes constant

$$
{}^{c} \text{caso}_{4}^{\text{o}} = \frac{\text{Ksp}}{\text{K}_{\text{[Cas}\,0_{\lambda}^{\text{o}}]}} \quad . \tag{36}
$$

Dissolution or precipitation of lime. An equation relating to the dissociation of lime in water with its constituent is shown as

$$
CaCO_{3} \div Ca^{++} + CO_{3}^{+}
$$
 [37]

and the solubility of calcite is usually described by the solubility product constant Ksp;

$$
Ksp = {}^{a}Ca \t{ }^{a}CO_{3}
$$
 [38]

where a is the activity of ions designated by the subscript. Since CO_3^- concentration is a function of partial pressure, and HCO₃ concentration is usually the predominant form in which $CO₂$ occurs in soil **water systems, it is more convenient to consider the following reactions.**

$$
H_2CO_3 + CaCO_3 \stackrel{+}{\leftarrow} Ca^+ + 2HCO_3^-
$$
 [39]

$$
K = \frac{a_{Ca} a_{HCO_3}^2}{a_{H_2CO_3}} \tag{40}
$$

If an equilibrium system is under constant pressure of $CO₂$ and if the activity coefficient of non-charged species (H_2CO_3) is unity, equation [40] becomes

$$
z = K^{c}H_{2}CO_{3} = {^{a}Ca} {^{a}H_{1}CO_{3}}
$$
 (41)

$$
ZE = \frac{Z}{\gamma_{Ca} \gamma_{HCO_3}^2} = {c_{Ca} \ c_{HCO_3}^2}
$$
 [42]

where γ is the activity of associate ion.

It has been pointed out by Olsen and Watanbe (1959) that the solubility of $CaCO₃$ in the soil is different from pure calcite, and the $\text{H}_{2}^\text{CO}_{3}$ content in the soil solution is variable at different moisture **contents. This in turn means that the value of Z, in the soil, varies** with water content. A comparison of equation [38] with [41] and [42] shows ^Zand ZE to be equivalent to the solubility product constant. Dutt et al. (1971) determined the following relationship between Z and water **content.**

$$
\log Z = -1.68 \log W - 4.46
$$
 [43]

where W is water content by weight expressed as percent. The same relation is used in the present model. ZE is then estimated from equation [42].

Using the same argument as in the case of solubility of gypsum that x_2 is the moles per liter of Ca⁺⁺ that dissolves or precipitates, then the equilibrium concentration of Ca^{++} and HCO₃ is

$$
c_{\text{Ca}} = c_{\text{Ca}}^{\text{o}} + \mathbf{x}_2 \tag{44}
$$

$$
^{c}_{\text{HCO}_3} = {^{c}_{\text{HCO}_3} + 2x_2} \tag{45}
$$

Substituting equation [44] and [45] in equation [42] results in a cubic equation

$$
A x_2^3 + B x_2^2 + C x_2 + D = 0
$$
 [46]

where

A = 4
\nB = 4
$$
\left[c_{CA}^{0} + c_{HCO_3}^{0} \right]
$$

\nC = $\left[c_{HCO_3}^{02} + 4 c_{CA}^{0} c_{HCO_3}^{00} \right]$
\nD = $\left[c_{HCO_3}^{0} \cdot c_{Ca}^{0} - ZE \right]$.

Equation [46] can be solved for x_2 by Newton Raphson iteration method.

Cation exchange. An equation that describes Ca-Mg exchange is

$$
\frac{a_{\text{Ca}}}{a_{\text{Mg}}} = K_{\text{Ca}-\text{Mg}} \frac{E_{\text{Ca}}}{E_{\text{Mg}}} \tag{47}
$$

where K_{Ca-Mg} is the exchange coefficient for Ca and Mg. Let y moles of Mg ⁺⁺ per gm of soil go into solution or are adsorbed. Let the initial concentration of Ca^{++} and Mg^{++} be c_{Ca}^0 and c_{Mg}^0 moles/liter in the solution phase and E^{O}_{Ca} , E^{O}_{Mg} be moles/gm adsorbed on the soil matrix. The change in the relative composition of Ca^{++} and Mg⁺⁺ from the inter**action of solution and** adsorbed phase is then

$$
E_{Ca} = E_{Ca}^0 - y
$$
 [48]

$$
E_{Mg} = E_{Mg}^0 + y
$$
 [49]

$$
c_{Ca} = c_{Ca}^0 + \beta y
$$
 [50]

$$
c_{\text{Mg}} = c_{\text{Mg}}^{\circ} - \beta y \tag{51}
$$

where β is the ratio of gm of soil to liter of solution. Combining

equation $[48]$ to $[51]$ with equation $[47]$ results in quadratic expres**s i an**

$$
Ay^2 + By + C = 0
$$
 [52]

whe re

$$
A = \beta \left(1 - K_{Ca-Mg} \right)
$$

\n
$$
B = \beta \left(E_{Mg}^{o} + K_{Ca-Mg} E_{Ca}^{o} \right) + C_{Ca}^{o} + K_{Ca-Mg} C_{Mg}
$$

\n
$$
C = C_{Ca}^{o} E_{Mg}^{o} - K_{Ca-Mg} C_{Mg}^{o} E_{Mg}^{o}.
$$

Equation [52] can be solved for y.

Gapon's equation was used to describe the non-symmetrical exchange between Ca and Na

$$
\frac{\sqrt{a_{\text{Ca}}}}{a_{\text{Na}}} = K_{\text{Ca}-\text{Na}} \frac{E_{\text{Ca}}}{E_{\text{Na}}} \tag{53}
$$

Using the same reasoning for calculating the equilibrium concentration as in the case of Ca-Mg exchange, equation [53] reduces to

$$
Ay_1^4 + By_1^3 + Cy_1^2 + Dy_1 + E = 0
$$
 [54]

where y_1 is the change in concentration required to reach equilibrium **from initial concentration, and**

$$
A = -4 K_{Ca-Na} \beta^{2}
$$

$$
B = 4\beta \left[\gamma_{1/2} + 2 K_{Ca-Na}^{2} E_{Ca}^{0} \beta + K_{Ca-Na}^{2} C_{Na}^{0} \right]
$$

where $\gamma_{1/2}$ is the ratio of activity coefficent of Ca to Na,

$$
C = 4\gamma_{1/2} \left(c_{Ca}^{\circ} + E_{Na}^{\circ} \beta \right) - 4K_{Ca-Na}^{2} \beta E_{Ca}^{\circ} \left(\beta E_{Ca}^{\circ} + 2c_{Na}^{\circ} \right)
$$

$$
- K_{Ca-Na}^{2} c_{Na}^{\circ^{2}}
$$

$$
D = E_{Na}^{\circ} \gamma_{1/2} \left(4c_{Ca}^{\circ} + E_{Na} \beta \right) + 2K_{Ca-Na}^{2} E_{Ca}^{\circ} c_{Na}^{\circ} \left(2\beta E_{Na}^{\circ} + c_{Na}^{\circ} \right)
$$

$$
E = E_{Na}^{\circ 2} c_{Ca}^{\circ} \gamma_{1/2} - K_{Ca-Na}^{\circ 2} c_{Ra}^{\circ 2} E_{Ca}^{\circ}.
$$

22

Equations [30), [35), [46), [52), and [54] are used to calculate the **equilibrium concentrations.**

MATERIALS AND METHODS

The applicability of the theoretical model was tested in the laboratory as well as in the field experiment.

Laboratory Experiment

In the laboratory the experiment was done using a lucite column packed with air dry soil. The soil used in the experiment was obtained from the Hullinger Farm near Vernal, Utah. The farm is located west of the airport, about 1 and 1/2 mile south of 5th East and Main Streets.

Column set up

The column consisted of 12 stacked rings with an inner diameter of 10.4 cm and an outer diameter of 11.4 cm. The top ring was 8.5 cm high, whereas the other 11 rings were 5.1 em high. The rings were interlocked by a groove and projection arrangement coated with petroleum gel to prevent leakage of water from the column. The whole column was bolted together by three brass rods. The bottom ring had a plate at the bottom with an outlet at its center to collect the effluent. To avoid sealing the outlet with soil, it was covered with a screen and a filter paper.

Packing of the column

In order to avoid layering of soil in the column while packing, the following procedure was adopted: Two sieves of 4 mm and 2 mm size were placed one above the other at the top of the column. Air
dried, sieved soil is passed through the 4 mm and then onto the 2 mm sieve at such a rate that sieves were not blocked. The column was filled to a height of about 61 em. The soil was leveled at the surface by hand. The uniformity in the packing was checked with the density probe (Davidson et al., 1963).

Column experiment

Three different cases of initial and boundary conditions were considered.

Case $#1$. "Sprinkler irrigation" condition with a layer of salt at the soil surface. $CaCl_{2} \cdot 2H_{2}0$ salt was applied at the rate of (4547.2) kg/ha) before wetting with the irrigation water. The soil was wetted by simulated sprinkler irrigation (0.57 cm/hr). The soil was leached until the wetting front nearly reached the bottom of the column (35.6 hrs). The column was then segmented. Each segment was weighed to estimate the bulk density. Portions of the soil from each segment were used to extract soil solution and to determine the water content. Exchangeable cations were determined from the other portion of the soil left in each segment. The chemical composition of irrigation water and soil is reported in Tables 1 and 2, respectively.

Table 1. Chemical composition of irrigation water

Table 2. Initial conditions for column experiments

Case #2. "Rain" condition with a layer of salt at the soil surface. The salts applied were $CaCl₂ + 2H₂0$ (1993.6 kg/ha), MgCl₂ \cdot 6H₂O (3225.6 $kg/ha)$, and NaCl (1523.6 kg/ha). The chemical composition of the soil used in this case is given in Table 2. The soil was wetted with distilled water by simulated rain (0.58 cm/hr) for 37.6 hours at which time the wetting front was nearly at the bottom of the column. The column was then segmented and analysed by the same procedure as in case $#1$.

Case #3. "Irrigation-evaporation-irrigation" with a layer of salt at each irrigation. Before wetting \texttt{CaCl}_{2} \cdot $\texttt{2H}_{2}$ 0 salt was applied at the rate of 4547.2 kg/ha. The soil was wetted for 35.4 hours by simulated sprinkler irrigation (0.59 cm/hr). The same soil was used as in case $#2$. The irrigation water applied had the same composition as in case #1. A water table was created at the end of infiltration and evaporation was allowed @0.095 cm/hr for 75.8 hours. The chemical composition of water of the water table was the same as that in the irrigation water. Evaporation was then continued without a water table and with the bottom outlet plugged for another 196.8 hours @0.0304 em/hour . At the end of evaporation the soil was wetted by

simulated irrigation (0.78 cm/hr for 23.7 hours) with a layer of salt at the soil surface. The salt applied was NaCl (3427.2 kg/ha). During infiltration the bottom outlet was unplugged. At the end of the last irrigation, the column was segmented and analysed by the same procedure as in cases #1 and #2.

Physical properties of the soil

In order to test the applicability of the computer model it is required to have appropriate data of the hydraulic properties of the soil. These properties are hydraulic conductivity-water content (Figure 1) and pressure head-water content (Figure 2) relationships. The data reported by Andrade (1971) for the given soil was used in this study.

Chemical analysis

Electrical conductivity of the soil solution was measured with a Beckman Model RC-19 conductivity bridge using a 2 ml pipet cell with **a cell constant of one. Measurements were taken at room temperature** and corrected to 25 C.

Chloride concentration was determined potentiometrically using a silver billet electrode and a saturated calomel electrode in conjunction with a corning model 12 expanded scale pH meter.

The solution extract was diluted with lanthanum oxide in concentrated HCl and the concentration of calcium, magnesium, and sodium was analysed by atomic adsorption spectophotometer (Perkin Elmer $Model - 303$.

Figure 2. Hydraulic conductivity-water content relationship for Mesa sandy loam.

Bicarbonate ion concentration in the irrigation water was calculated using the relationship

$$
H_2O + CO_2 \stackrel{K}{\underset{\leftarrow}{\right}} H^+ + HCO_3^-
$$
 [55]

$$
K = \frac{a_{\text{H}} a_{\text{HCO}_3}}{c_{\text{CO}_2}}
$$
 [56]

where a is the activity of the ions designated, $a = m\gamma$, m is the molality and γ is the activity coefficient. At the reference state $a_{H_2^00} = 1$ in pure water. Thus, co_2 can be replaced by $m_{CO_2^0}$ = cP_{CO_2} , where c is Henry's law constant (0.0344 at 25 C), P_{CO_2} is partial pressure of CO_2 in atmosphere (3 x 10⁻⁴) and K = 4.45 x 10⁻⁷. **Bicarbonate ions concentration in the soil solution was estimated** from the relationship given by equation [38] and

$$
HCO_3 \stackrel{K}{\leftarrow} H^+ + CO_3^-
$$
 [57]

$$
K = \frac{4H}{a_{HCO_3}} \tag{58}
$$

where $K = 4.69 \times 10^{-11}$. Details of this method are given by Olsen and Watanbe (1959). These concentrations were just an approximation to start with and were corrected by subroutine EXCH to bring into equilibrium with the system.

The concentration of sulphate ions was estimated from the difference of total salts and the summation of chloride and bicarbonate **ions.**

For the analysis of exchangeable cations, 5 to 7 gms of wet soil sample was washed with 150 ml of 95 percent ethanol in a leaching funnel until free of chloride. Exchangeable calcium and magnesium were extracted by leaching with 100 ml of lN sodium acetate at pH 8.2. A similar procedure was followed for exchangeable sodium except that it was extracted with lN ammonium acetate (pH 7.0). Analysis of cations was made as described previously.

Because of the insufficient amount of solution, collected in the field experiments for chemical analysis, soil samples were collected and saturation paste prepared (Richards, 1954). The saturation extracts were analysed as described previously.

Exchange constant

Exchange constants are defined by equation $[47]$ and $[53]$. Their values were determined from the known concentration of ions in the solution and exchangeable phase. Although their names implies a constant value, they vary with the total salt concentration in the present system. It would be more appropriate to define them as exchange coefficient rather than exchange constant. Figures 3 and 4 are the plot of these coefficients with total salt concentration. Exchange coefficients for case #1 are given by Figures 3 a and 4a, while for case #2 and #3. their values are given in Figures 3b and 4b.

Field Experiment

The field experiment was conducted on the Hullinger Farm near Vernal, Utah. The soil type was Mesa sandy loam. Tensiometer, salinity sensors, and four probe units were installed in the center of the plot

Figure 3. Variation of Ca-Mg exchange coefficient with solution concentration a) surface soil, b) sub soil.

Figure 4. Variation of Ca-Na exchange coefficient with solution concentration a) surface soil, b) sub soil.

at 15, 45, 75, 105, and 165 cm depth at site A, and, 15, 45, 75, 135, and 165 cm depths at site B. Duplicate tensiometer cups were also installed at the given depths to obtain soil solution samples. To facilitate the collection of reasonable amounts of soil solution for electrical conductivity (EC) measurements , suction was applied on the solution cups for 8 to 12 hours depending upon the moisture content of the soil. Soil samples were collected for chemical analyses at three different times during the experiment. Samples were taken at 30 em intervals to a depth of 120 em. Alfalfa was the major crop grown. Sprinklers were used as means of irrigation. $CaCl₂ + 2H₂0$ (4390.4) kg/ha) and NaCl (3561.6 kg/ha) were applied before the first and second irrigation cycles. Initial and boundary conditions to the above experiment are given in Tables 3 and 4, respectively. Chemical composition of irrigation water was reported in Table 1. Soil moisture distribution was determined by the neutron probe and water loss by evapotranspiration was estimated from the lysimeter data.

Computer Model

The computer model of Dutt et al. (1971) for solubility of minerals and exchange between ions in solution and soil was combined with the salt and water transport model of Bresler and Hanks (1969). The resultant model consists of a main program and five sub programs . The sub programs are designated as: (1) PLOT, (2) EXCH, (3) EQEXCH, (4) SALT, and (5) ACOF.

Main program

Figure 5 is the flow chart for the main program. The main program has several major responsibilities. First it reads the initial and

| Depth | Water Content | Calcium | Magnesium | Sodium | Sulphate | Chloride | Bicarbonate | Gypsum |
|-------------|------------------|---------|-----------|--------|----------|----------|-------------|-------------|
| (cm) | (θ) | (me/1) | (me/1) | (me/1) | (me/1) | (me/1) | (me/1) | (gm/100 gm) |
| $0 -$ 30 | 0.2347 | 36.0 | 14.31 | 5.04 | 44.25 | 10.35 | 0.75 | 0.0 |
| $30 - 55$ | 0.2446 | 25.75 | 14.19 | 4.22 | 34.08 | 9.16 | 0.92 | 0.5 |
| $55 - 100$ | 0.2764 | 27.25 | 16.29 | 7.13 | 44.05 | 5.67 | 0.95 | 0.5 |
| $100 - 115$ | 0.3053 | 35.10 | 19.95 | 4.91 | 57.84 | 1.30 | 0.82 | 0.5 |
| $115 - 165$ | 0.3661 | 32.0 | 16.25 | 4.39 | 50.54 | 1.26 | 0.83 | 0.5 |

Table 3. Initial conditions for field experiments

| Time (hrs) | Flux at the Surface $(10^2 \times cm/hr)$ | ET Flux $(10^1 \times cm/hr)$ | Comment |
|---------------|--|----------------------------------|------------|
| | | | |
| 24.0 | -0.300 | -0.300 | |
| 16.0 | 0.0064 | 0.0 | Irrigation |
| 8.0 | 0.00 | 0.00 | |
| 24.0 | -0.211 | -0.211 | |
| 120.0 | -0.187 | -0.187 | |
| 72.0 | -0.135 | -0.135 | |
| 60.0 | -0.219 | -0.219 | |
| 15.0 | 0.0062 | 0.0 | Irrigation |
| 69.0 | -0.224 | -0.224 | |
| 2.0 | 0.0037 | 0.0 | Rain |
| 70.0 | -0.190 | -0.190 | |
| 24.0 | 0.0021 | 0.0 | Rain |
| 24.0 | -0.226 | -0.226 | |
| 24.0 | 0.005 | 0.0 | Rain |
| 24.0 | -0.227 | -0.227 | |
| 24.0 | -0.198 | -0.198 | |
| 27.0 | -0.191 | -01.91 | |

Table 4. Boundary condition in the field experiment for soil water flow

Figure 5. Flow diagram for MAIN.

boundary conditions. The initial conditions in this experiment include the concentration of salts applied on the soil surface at the beginning of the experiment. Since the salts were applied in the powder form rather than in solution, it was assumed that they were soluble at the given water content of the soil. The program then calls for subroutine EQEXCH. This subroutine calculates $\text{Cas}\, 0_A$, MgSO_{Λ} ion-pairs and equilibrium concentration of exchangeable ions. The input and transformed data are then printed to provide the user with a record. Concurrently , the main program calls for subroutine PLOT which plots the water and salt content with depth. After setting various counters and initializing certain values the program computes the new values of pressure **head and water content.**

The routine then executes a large outer loop for the number of depths in a profile. Within this loop the routine checks for the amount of water leaving or entering the top or bottom of a soil segment at a particular depth. If the amount of moisture flow is not negligible subroutine SALT is called which computes the flow of salts due to mass flow of water. A check is then made to call the subroutine EXCH. The check insures that changes in concentration of ions due to solubility of minerals and exchange are calculated every hour rather than every Δt . No great difference was noted in the predicted values when subroutine EXCH was called each Δt . The counter for this check is initialized to zero after each call for subroutine EXCH.

Then the program increments the time counter with Δt and initializes the old values with the recently computed values of the variables used in salt and water flow. The routine then calls for subroutine PLOT and prints the output.

A check is then made for the new boundary conditions and cumulative time for which the program is allowed to run. If the time equals the cumulative time then it stops after the subroutine PLOT is called and the needed output information is printed. Otherwise, it goes back to statement 16 and executes for the next Δt increment.

Input data. The input data needed are as follows.

1. Hydraulic conductivity-water content and pressure head-water content tabular data covering the range of water content to be encountered during the period of interest (basic soil property).

2. Air dry and saturated soil water contents (basic soil data) .

3. Root distribution with depth (active roots for adsorbing water) for the period. At present the model has no provisions for changing this with time (basic plant property) .

4. Plant water potential below which the plant wilts and the actual transpiration will be less than potential transpiration (basic plant property).

5. Activity coefficient-ionic strength tabular data covering the range of ionic strength encountered in the system.

6. Water content-depth tabular data at the beginning (initial conditions).

7. Chemical composition depth tabular data at the beginning (initial conditions). This involves the knowledge of the chemical analysis of the important chemical species . At present it considers Ca, Mg, Na cations and C1, SO_4 , and HCO₃ anions.

8. Potential transpiration and potential evaporation rate or portential irrigation or rainfall rate as a function of time for the period (boundary conditions).

9. Chemical composition of the irrigation or rain water (boundary conditions).

10. Presence or absence of a water table at the bottom of the soil (boundary condition).

Output data. The type of output data that is available is almost infinite. Consequently, a selection of the desired data is made from the following.

1. Soil water content and pressure head *VB* depth and time during the period.

2. Chemical composition of the soil solution *VB* depth and time during the period.

3. Estimated evaporation and transpiration as a function of time.

4. Water flow into the water table or up from the water table as a **function of time.**

5. Chemical composition of the water going into the water table or up from the water table as a function of time.

6. Estimated plant water potential as a function of time.

Subroutine PLOT

This subroutine plots the salt and water content with depth.

Subroutine EXCH

Figure 6 is the flow chart for this subroutine. The subroutine is called in the main program approximately every hour or at each Δt if Δt is greater or equal to one hour. This implies that in a time of one hour equilibrium is established between the ions in the solution and solid phase. Adjustment in the concentration of different ions due

Figure 6. Flow diagram for EXCH.

to solubility of minerals and exchange with soil particles is made in this subroutine. The adjusted concentrations are then returned to the **main program.**

Subroutine EQEXCH

Since the concentration of exchangeable cations are necessary to predict changes in soil solute composition, and, reliable analytical methods are not available when excess calcium carbonate is present. an improved method for their calculation is necessary. This subroutine calculates exchangeable ions from initial soil analysis. It also calculates the concentration of Ca^{++} , Mg^{++} , $SO_L^-,$ CaSO_{Λ} ion-pair and $MgSO_4$ ion-pair from their total analysis. Theory underlying this subroutine is discussed in the following sections.

Sulphate occurs in basic solution in more than one form. In addition to free sulphate ion, there are two forms which have been shown to be of importance in base saturated soil-water systems; these are undissociated, soluble CaSO₄ (Dutt, 1964) and MgSO₄ (Tanji and Doneen, 1966). Thus the total sulphate in solution is

$$
{}^{c}TSO_{4} = {}^{c}SO_{4} + {}^{c}CasO_{4} + {}^{c}MgSO_{4}. \qquad [59]
$$

Similarly, the total calcium, C_{TCa} and magnesium C_{TMg} is

$$
{}^{c}TCa = {}^{c}Ca + {}^{c}Cas0,
$$
 [60]

$$
{}^{c}TMg = {}^{c}Mg + {}^{c}MgSO_{\Lambda}.
$$

The thermodynamic equilibrium constant for equilibrium between the undissociated species in solution and the appropriate ions is

$$
K_{\text{[Cas0]}_{4}^{o}} = \frac{a_{\text{Ca}} a_{\text{SO}_{4}}}{a_{\text{Cas0}^{o}}}
$$
 (62)

and

$$
K_{[MgSO_4^0]} = \frac{a_{Mg} a_{SO_4^0}}{a_{MgSO_4^0}}.
$$
 (63)

Combining equation [60) and [62), we get

$$
{}^{c}c_{a50}{}^{o}_{4} = \frac{{}^{Y}s_{0} {}^{c}_{4} {}^{Y}c_{a} {}^{c}r_{Ca} {}^{c}s_{0} {}^{c}_{4}}{K_{[cas0}{}^{o}_{4}{}^{1} + {}^{Y}c_{a} {}^{Y}s_{0} {}^{c}_{4} {}^{c}s_{0} {}^{c}_{4}} \qquad (64)
$$

and similarly combining equation [61) with [63) we get

$$
c_{\rm MgSO_4^o} = \frac{\gamma_{\rm SO_4} \gamma_{\rm Mg} c_{\rm TMg} c_{\rm SO_4}}{K_{\rm [MgSO_4^o]}} + \gamma_{\rm Mg} \gamma_{\rm SO_4} c_{\rm SO_4} \tag{65}
$$

Combining equation [59) with [64) and [65] result in a cubic **equation**

$$
A x3 + B x2 + C x + D = 0
$$
 [66]

where

$$
x = {}^{c}S0_{4}
$$
\n
$$
A = \gamma_{2}^{4} = [\gamma_{Ca} \gamma_{SO_{4}}]^{2} = [\gamma_{Mg} \gamma_{SO_{4}}]^{2}
$$
\n
$$
B = \gamma_{2}^{2} \Big((K_{[CaSO_{4}^{O}]1} + K_{[MgSO_{4}^{O}]1}) + \gamma_{2}^{2} (c_{TMg} + c_{TMg} - c_{TSO_{4}}) \Big)
$$
\n
$$
C = K_{[CaSO_{4}^{O}]1} K_{[MgSO_{4}^{O}]1} + \gamma_{2}^{2} \Big[c_{TMg} K_{[CaSO_{4}^{O}]1} + c_{TCa} K_{[MgSO_{4}^{O}]1}
$$
\n
$$
- c_{TSO_{4}} K_{[CaSO_{4}^{O}]1} + K_{[MgSO_{4}^{O}]1}
$$
\n
$$
D = c_{TSO_{4}} K_{[MgSO_{4}^{O}]1} K_{[CaSO_{4}^{O}]1}.
$$

Equation [66] is solved by Newton Raphson method for c_{S0} . 4 Concentration of Ca^{++} , Mg⁺⁺ are calculated using equations [60], [61], [64], and [65].

Equations [47] and [53] are used to describe Ca-Mg and Ca-Na exchange in this model. If calcium, magnesium, and sodium are the only cations in the soil then

$$
E_T = E_{Ca} + E_{Mg} + E_{Na}
$$
 [67]

where E_T is the cation exchange capacity. Combining equations [47], and [53] with [67] results in

$$
E_{Ca} = \frac{E_{T}}{\left[\frac{a_{Na}}{K_{Na}-Ca} \frac{K_{Ca}-Mg}{a_{Ca}} \frac{a_{Mg}}{a_{Ca}} + 1\right]}
$$
 (68)

Exchangeable cation concentrations are calculated using equations [53], [67], and [68].

Figure 7 is the flow chart for this subroutine . This subroutine is called only once at the start in the main program.

Subroutine SALT

The subroutine calculates the changes in salt concentration due to mass flow of water. The mass flow of salt is computed by equation [23].

Subroutine ACOF

This subroutine is called in both subroutines EQEXCH and EXCH. It calculates the activity coefficients of monovalents and divalent ions. In the dilute solution activity coefficient of ions can be adequately described by a modified form of Debye Hiickel law for mixed electrolytes

Enter moles/ $i \leftarrow me/l$ K_{Na} . Ca K_{Ca-Mg} E_T CALL ACOF $C_{SO4} + Eq. 66$ $\overline{\text{CaSO}_4 \leftarrow \text{Eq }64}$ $MgSO₄$ \leftarrow Eq 65 ϵ_{Ca} Eq 68 $\ddot{}$ E_{Na} \leftarrow Eq 53 Eq 67 $\boldsymbol{\epsilon}_{\mathsf{Mg}}$ \leftarrow me/l \leftarrow moles/l Return

Figure 7. Flow diagram for EQEXCH.

$$
\log \gamma_{i} = -\frac{0.509 \, z_{i}^{2} \sqrt{I}}{1 + \sqrt{I}}
$$
 [69]

where

$$
I = \frac{1}{\rho_0} \frac{1}{2} \frac{n}{1} \sum_{i=1}^{n} c_i Z_i
$$
 [70]

where i is the ion species of interest, n is the total number of ion species in solution, Z is the valence, and ρ is density of the solution $(\rho_{0} = 1).$

In concentrated salt solution as used in the present experiments, table values (Robinson and Stokes, 1955) of activity coefficients are used. Activity coefficients of individual ions are calculated from the mean molal activity coefficient of salts using the following relationship

$$
\gamma_{\text{Ca}} = \frac{\gamma_{\pm}^3 \text{ CaCl}_2}{\gamma_{\pm}^2 \text{KCl}} \tag{71}
$$

$$
\gamma_{\text{Na}} = \frac{\gamma_{\pm}^2 \text{ NaCl}}{\gamma_{\pm} \text{KCl}} \ . \tag{72}
$$

As a first approximation, it is assumed

$$
\gamma_{Ca}^{++} = \gamma_{Mg}^{++} = \gamma_{SO}^{--}
$$
 [73]

$$
\gamma_{\text{Na}}^{+} = \gamma_{\text{HCO}_3}^{-} \tag{74}
$$

This subroutine looks up the table value of divalent or monovalent ion activity coefficient corresponding to a given ionic (I) strength .

RESULTS AND DISCUSSION

To test the applicability of the model predicted values are **compared with the experimental measurements. Results of each case are** discussed separately.

Column Experiment

Case $#1$

Figure 8a is the plot of experimental and predicted soil moisture distribution after 35.6 hours of infiltration. Total salt distribution curves corresponding to the soil moisture distribution are given in Figure 8b. Figure 8a shows that the predicted water content is slightly higher than the measured values at depths greater than 40 em and vice versa at depths below 40 em. These differences are due to the hydraulic parameters used in this model which were determined for an undisturbed soil while in the column the soil was loosely packed.

Observed and predicted total salt concentration (Figure 8b) have the same distribution pattern. The depth at which the maximum concentration occurs is almost the same in both cases. The concentration of salt in the upper 18 cm had the same total concentration as that of the irrigation water. However, below 18 em there are some differences between the measured and predicted values. The predicted concentrations are less than the measured values between 18 and 33 em while below 33 em, the reverse is true. These differences are hard to explain with the present state of information. One of the reasons for high predicted salt concentrations below 40 cm is the low predicted water content. However, the agreement is considered good enough for most purposes.

Figure 8. Comparison of predicted and measured a) water content, b) total salt concentration for condition of case #1.

The comparison of the experimental and predicted concentrations of individual ions comprising the total salts is plotted in Figure 9. Since the sulphate and bicarbonate ion concentrations were calculated and not measured the plot for these ions are not drawn. Calcium and magnesium includes the corresponding ions and ion-pairs. Ion-pairs of sodium ions are assumed to be the same as the total sodium concentration.

In general both the cations and anions followed the same distribution as that of total salts. The concentration of each ion increased with depth and maximum concentration occurred at about 58 em. Predicted concentration of calcium, magnesium, and sodium seemed to be in close correspondence with the experimental values at low salt concentration. However, there is a significant difference between computed **and measured cation concentrations at total salt concentration greater** than 45 me/1. This lack of agreement seems to result from the inadequate description of the cation exchange process at the higher salt concentrations. Since measured and predicted total salts distribution are in reasonable agreement, it is expected that the cations comprising the total salts also follows the same trend. This, however, is not ture. The predicted calcium concentration is about 1 and 1/2 to 2 times greater than its measured concentration below 30 em depth while the predicted magnesium and sodium concentrations are less than the measured values. Since relative concentration of each cation is controlled by the exchange coefficient, the above differences seems to be due to inadequate information concerning exchange coefficients.

Figure 9. Comparison of predicted and measured ion concentration profiles for conditions of case $#1$.

It has been assumed in this study that a given exchange coefficienttotal salt concentration relationship holds at all levels of salt concentration. This assumption is not necessarily true as discussed in the previous paragraph. Since concentration of each cation at the end of infiltration is affected by the composition of solution in the early hours of infiltration, the correct values of exchange coefficients seems to be important at high salt concentrations in soils.

Another cause for disagreement may be the assumption involving the activity coefficients. It has been assumed that $\gamma_{Ca} = \gamma_{Mg} = \gamma_{SO_4}$ Table values (Robinson and Stokes, 1955) indicate $\gamma_{\rm SO_4}$ < $\gamma_{\rm Ca}$ < $\gamma_{\rm Mg}$ in pure solution. The difference between their values is quite prominent at high ionic strength.

Still another cause for disagreement may be that Na_2SO_4 , CaCO_3 , $MgCO₃$, and other complex ion formations are not considered. This may be important because there is an increase in complex ion formation with increase in salt concentration.

Case $#2$

Three different salts (CaCl₂, MgCl₂, and NaCl) having a common anion were leached with distilled water. Figure 10 shows the water content and total salt distribution at the end of 37.6 hours of infiltration. There is a good agreement between the observed and the predicted water content distributions. However, the measured total salt concentration is quite different than the predicted values. There is not any regular increase with depth of the measured total salts. There is no well defined depth at which maximum concentrations of total salts occurs. The zig-zag distribution of total salts indicates that there

Figure 10. Comparison of predicted and measured a) water content and b) total salt concentration profiles for conditions of case #2.

may have been alternate addition of salty and non-salty water at the surface. It is postulated that this kind of behavior results from the assumption that all the salts are soluble at initial soil water content. Since the salts were applied in the powder form, it seems that some of the salts were not solublized after the first wetting. Because the water was applied in drops and manually checked for its uniform application at the surface, it is believed that these salts eventually became dissolved at irregular time intervals and led to this type of distribution. There may also be analysis problems or errors unaccounted for.

Figure 11 gives the distribution of individual ions. Measured concentration of almost all the ions follows the same general distribution of total salts. There is a poor agreement between the measured and computed values for all the ions. The model at its present stage does not predict this kind of distribution.

Case $#3$

This is the case where wetting-drying-wetting cycle was followed. Figure 12 shows the soil moisture and total salt distribution at the end of the experiment. Agreement between measured and predicted water content is reasonable considering the assumption that hydraulic properties used were for an undisturbed sample. Measured total salts distribution indicated the presence of two peaks at about 12 and 52 em depths, while the model predicted a single depth (38 em) at which maximum concentration occurred. Since enough water was applied at the second irrigation, it is expected that all the salts would have moved to the bottom of the column at the end of the second infiltration. As sodium ions (Figure 13c) are the major component contributing to the peak in

Figure 12. Comparison of measured and predicted a) water content and b) total salt concentration profiles of case #3.

total salt concentration, at 12 em depth, it seems that all the NaCl salt added before the second irrigation apparently did not dissolve immediately after irrigation, as is assumed in the model. Crystals left undissolved eventually became solublized at some later time in the irrigation cycle and led to the peak in total salt and other ions at the 12 em depth.

Figure 13, the plot of individual ions, shows the same distribution pattern as that of total salts. Measured chloride ion distribution also supports the postulation discussed in the last paragraph.

Field Experiment

The model was tested under field conditions at Hullinger Farm near **Vernal, Utah. Water movement and, thus, the salt movement due to the** presence of roots was also considered. In order to avoid the complexity arising due to layered soil, the soil profile was assumed to have uniform properties throughout. Presence of gypsum was considered in the initial conditions below 30 em depth. The model was tested over a period of two drying and wetting cycles. Hysteresis in the hydraulic properties was ignored. Comparison of the predicted and measured values was made at three different times in the cycle. Since no measurements of individual ions were made on the solution samples **at field water content, an approximate method was used to arrive at the** concentrations from the saturation extraction analysis. The method involves the assumption that the individual ion concentration changes in the same proportion with changing water content as does the electrical conductivity of the solution. This assumption may not be exactly valid for complex ions and ions which react with the soils. Since chloride

ions do not interact with the soil, it is expected that the given assumption holds good for chloride.

Figure 14 is the plot of water content distribution at three different times during the cycle. There is a good correspondence between the measured and predicted values on the first (Figure 14a) and the third (Figure 14c) samplings. Although, the predicted water content does not agree too well with the measured values on the second sampling (Figure 14b) both distribution follows the same trend.

Electrical conductivities of solution at field water content and saturation extract are plotted in Figure 15. The ratio of the two **values was used to correct the saturation extract analysis to get ion** concentration at field water content. Except the first sampling (Figure 16a), the predicted values closely relate the corrected total salt concentration. Since the saturation extract analysis represented **an average of 30 em depth, the corrected concentrations are represented** by histograms. Depth at which the maximum concentration occurs is deeper in measured than in predicted distribution. This may be the result of discontinuity in the measured distribution.

Figures 17, 18, and 19 give the individual cation distribution. Measured calcium concentrations are generally less than the predicted **values, while the reverse is true for sodium and magnesium concentra**tions. These differences seem to result from the assumption that cation concentration changes in the same proportion as does the EC of the solution. Since the preference of exchanger for the ions of higher valence increased with dilution of the solution (Helfferich, 1962) it is expected that the proportionate increase in calcium and magnesium be more than in sodium. The approximation used to get the ion concentration

Water content (θ)

Figure 14. Comparison of predicted and measured water content profiles for alfalfa, a) 324 hrs, b) 339 hrs, and c) 627 hrs.

Electrical conductivity (m mhos/cm)

Figure 15. Comparison of saturation extract and soil solution electrical conductivity profiles for alfalfa, a) 324 hrs, b) 339 hrs, and c) 627 hrs.

Total salt concentration $(me/1)$

Figure 16. Comparison of predicted and measured total salt concentration profiles for alfalfa, a) 324 hrs, b) 339 hrs, and c) 627 hrs. Both the predicted histogram and continuous **curves are shown.**

Calcium (me/1)

Figure 17. Comparison of predicted and measured calcium concentrations profiles for alfalfa, a) 324 hrs, b) 339 hrs, and c) 627 hrs. Both the predicted histogram and continuous **curves are shown.**

Magnesium $(me/1)$

Figure 18. Comparison of predicted and measured magnesium concentration profiles for alfalfa, a) 324 hrs, b) 339 hrs, and c) 627 hrs. Both the predicted and histogram and continuous curves are shown.

Sodium $(me/1)$

 $\mathcal{L}_{\mathbf{A}}$

Figure 19. Comparison of predicted and measured sodium concentration profiles for alfalfa, a) 324 hrs, b) 339 hrs, and c) 627 hrs. Both the predicted and histogram and continuous curves are shown.

at field water content, however, assumes the same dilution effect for all the cations.

Sodium concentration measured in the saturation extract is also higher than the computed values while the opposite is true for the **magnesium and calcium concentrations. Since the area under total salt** curves is about the same for both measured and predicted distribution, the relative concentration of cations depends upon the exchange coefficients as discussed previously in case #1.

Chloride concentrations are plotted in Figure 20. It follows the same distribution as that of total salts. The agreement between the measured and predicted chloride concentration is fair. However, the depth at which the maximum chloride ion concentration occurs is different in both measured and predicted distribution. This lack of **agreement seems to be due to the discontinuous nature of measured** chloride distribution curve.

Figure 21 gives the depth and salt concentration of the drainage **water during this experiment. Since no measurements were made only** computed values are plotted. It shows the capability of this model to provide this kind of information, which is useful in devising a scheme for quality control of irrigation return flow.

Chloride (me/1)

Figure 20. Comparison of predicted and measured chloride concentration profiles for alfalfa, a) 324 hrs, b) 339 hrs, and c) 627 hrs. Both the predicted histogram and continuous curves are $\sum_{i=1}^{\infty}$

Figure 21. Predicted drainage and salt concentration of the drainage water over a period of 627 hrs.

SUMMARY AND CONCLUSIONS

A model was developed to describe the simultaneous flow of water and salts in soils under varying initial and boundary conditions. Water and thus the salt movement due to plant roots extraction was also considered. To predict the distribution of adsorbed ionic species, correction due to sink or source term was made in their concentration. Specific **chemical processes contributing to sink or source term in the model are.:**

1. Dissolution or precipitation of gypsum and lime.

2. Formation of undissociated Ca and Mg sulphate.

3. Exchange between cations in solution and the soil matrix.

The principles of solubility product and equilibrium exchange were used. The solutes considered were Ca^{++} , Mg^{++} , Na^+ , and Cl^- . The model was tested under field and laboratory conditions. In the laboratory three cases with different initial and boundary conditions were studied. In the field the experiment was conducted with alfalfa as the major crop.

Tensiometer, salinity sensors, four probe units, and solution cups were installed in the center of the plots. Two wetting and drying cycles were followed. Measurements of water content and salt concentrations were made three times in the experiment. Experimental measurements were then compared with the predicted values.

There was a close correspondence between the measured and predicted water content in all experiments. However, predicted total salt concentration agreed fairly well with the measured values only in the

field and one of the column experiments. Chloride ions followed the same distribution pattern as that of total salts in the above experiments. Predicted calcium concentration was higher than the measured values while the opposite was true for predicted magnesium and sodium concentration. It is postulated that these differences results because of (1) insufficient description of the exchange and activity coefficients at high salt concentrations, (2) and other complex ion formation not included in the model at the present time.

In two laboratory experiments there was a poor agreement between the predicted and measured total salt concentration. The lack of agreement seems to result from the assumption involved in the present model or some reasons unknown at the present time. Since the salts were applied in the powder form they were assumed to be soluble at the initial water content. It was concluded that this assumption is one of the reasons for the apparent differences in the predicted and measured values. Different ion concentrations in these experiments followed the same distribution as the total salts.

The investigation regarding the applicability of model suggests that more tests are needed. It does appear to yield approximately correct values for total salt but individual species are not as accurately described.

SUGGESTIONS FOR FURTHER RESEARCH

Probably no research has ever been conducted that did not bring **up more questions than were answered. This one is no exception. Since** the model was not tested under a variety of initial and boundary conditions for salt flow, it should be further investigated in the field and laboratory to determine its suitability. In addition, there appears to be at least six related areas where more investigation is needed to improve and test this model. They are:

1. What is the effect of "built in" dispersion in numerical methods on the salt flow?

2. How do the activity coefficients of different ionic species vary at high salt concentration in soil solutions?

3. How do the exchange coefficients behave in the mixed salt solutions at high salt concentrations?

4. What is the correction due to other complex ion formations at high salt concentrations?

5. Because it was felt in this study that powdered salt does not dissolve immediately after wetting under unsaturated flow conditions, it is suggested that salt solutions rather than salt crystals should be used. This may be a problem in the field and needs further **consideration.**

6. The extraction of a sufficient amount of solution for chemical analysis, under field condition was difficult. Further research on the methods of solution extraction is desirable.

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APPENDICES

Appendix A. Tables

| Depth | Bulk Density | Water Content | $\mathop{\text{EC}}$ | Total Salts | Calcium | Magnesium | Sodium | Chloride |
|---------------|-----------------|------------------|----------------------|----------------|---------|-----------|--------|----------|
| (cm) | | (θ) | $(\mu$ mhos) | (me/1) | (me/1) | (me/1) | (me/1) | (me/1) |
| $0 - 5.1$ | 1.15 | 0.346 | 866 | 9.3 | 4.4 | 2.9 | 2.0 | 0.6 |
| $5.1 - 10.2$ | 1.16 | 0.353 | 744 | 8.5 | 4.3 | 2.6 | 1.7 | 0.3 |
| $10.2 - 15.2$ | 1.20 | 0.353 | 852 | 11.9 | 4.4 | 4.4 | 3.2 | 0.3 |
| $15.2 - 20.3$ | 1.19 | 0.350 | 1100 | 12.4 | 4.6 | 6.0 | 1.7 | 1.5 |
| $20.3 - 25.4$ | 1.19 | 0.348 | 2410 | 24.4 | 9.0 | 13.1 | 2.3 | 16.1 |
| $25.4 - 30.5$ | 1.21 | 0.340 | 3162 | 29.8 | 10.9 | 16.5 | 2.5 | 23.6 |
| $30.5 - 35.6$ | 1.17 | 0.331 | 3448 | 38.8 | 14.1 | 21.8 | 2.9 | 24.6 |
| $35.6 - 40.6$ | 1.14 | 0.332 | 5440 | 58.8 | 21.0 | 34.4 | 3.4 | 46.7 |
| $40.6 - 45.7$ | 1.17 | 0.326 | 6000 | 79.1 | 27.5 | 46.9 | 4.7 | 52.9 |
| $45.7 - 50.8$ | 1.14 | 0.322 | 7190 | 94.2 | 31.5 | 56.3 | 6.5 | 64.3 |
| $50.8 - 55.9$ | 1.15 | 0.300 | 6524 | 90.2 | 30.6 | 53.8 | 5.8 | 52.6 |
| $55.9 - 61.0$ | 1.15 | 0.281 | 13840 | 167.3 | 41.9 | 114.9 | 10.5 | 144.0 |

Table 5. Measured bulk density, water content, electrical conductivity, and ion concentration profiles for conditions of case $#1$

| Depth | Bulk Density | Water Content | EC | Total Salts | Calcium | Magnesium | Sodium | Chloride |
|---------------|------------------------|------------------|---------------------|----------------|---------|-----------|--------|----------|
| (cm) | | (θ) | (μmhos) | (me/1) | (me/1) | (me/1) | (me/1) | (me/1) |
| $0 - 5.1$ | 1.10 | 0.397 | 379 | 4.4 | 1.6 | 0.6 | 2.2 | 0.2 |
| $5.1 - 10.2$ | 1.12 | 0.386 | 552 | 6.4 | 2.6 | 1.1 | 2.7 | 0.3 |
| $10.2 - 15.2$ | 1.15 | 0.405 | 1894 | 26.2 | 12.3 | 8.9 | 5.0 | 7.3 |
| $15.2 - 20.3$ | 1.15 | 0.40 | 6612 | 84.3 | 36.3 | 26.7 | 11.4 | 54.3 |
| $20.3 - 25.4$ | 1.15 | 0.377 | 8064 | 98.2 | 45.0 | 38.3 | 14.9 | 82.5 |
| $25.4 - 30.5$ | 1.10 | 0.267 | 6480 | 61.2 | 29.4 | 23.5 | 8.3 | 47.5 |
| $30.5 - 35.6$ | 1.22 | 0.396 | 4836 | 43.6 | 20.6 | 16.4 | 6.6 | 22.0 |
| $35.6 - 40.6$ | 1.18 | 0.397 | 12031 | 142.2 | 61.9 | 52.1 | 28.3 | 87.8 |
| $40.6 - 45.7$ | 1.15 | 0.385 | 8252 | 75.9 | 39.4 | 26.7 | 9.9 | 68.0 |
| $45.7 - 50.8$ | 1.12 | 0.370 | 10186 | 125.0 | 62.5 | 54.2 | 8.3 | 104.0 |
| $50.8 - 55.9$ | 1.13 | 0.359 | 9590 | 90.1 | 47.4 | 35.4 | 7.3 | 86.5 |
| $55.9 - 61.0$ | 1.13 | 0.324 | 9230 | 115.3 | 59.1 | 49.6 | 6.6 | 81.8 |

Table 6. Measured bulk density, water content, electrical conductivity, and ion concentration profiles for conditions of case $#2$

a-

| Depth | Bulk Density | Water Content | $\mathbb{E}\mathbb{C}$ | Total Salts | Calcium | Magnesium | Sodium | Chloride |
|---------------|------------------------|------------------|------------------------|----------------|---------|-----------|--------|----------|
| (cm) | | (θ) | $(\mu$ mhos) | (me/1) | (me/1) | (me/1) | (me/1) | (me/1) |
| $0 -$ 5.1 | 1.11 | 0.376 | 1408 | 20.8 | 12.2 | 4.0 | 4.6 | 2.9 |
| $5.1 - 10.2$ | 1.13 | 0.388 | 6090 | 77.0 | 16.5 | 5.7 | 54.8 | 33.5 |
| $10.2 - 15.2$ | 1.12 | 0.376 | 11040 | 128.2 | 50.0 | 36.9 | 41.3 | 100.4 |
| $15.2 - 20.3$ | 1.16 | 0.394 | 7372 | 76.5 | 38.5 | 20.0 | 18.0 | 52.2 |
| $20.3 - 25.4$ | 1.16 | 0.412 | 6532 | 78.0 | 38.0 | 22.4 | 17.6 | 43.9 |
| $25.4 - 30.5$ | 1.16 | 0.404 | 5745 | 63.2 | 29.3 | 20.2 | 13.7 | 34.5 |
| $30.5 - 35.6$ | 1.18 | 0.416 | 5320 | 61.4 | 29.5 | 17.1 | 14.8 | 29.0 |
| $35.6 - 40.6$ | 1.16 | 0.419 | 6128 | 70.9 | 34.5 | 22.4 | 14.1 | 35.7 |
| $40.6 - 45.7$ | 1.14 | 0.418 | 7380 | 83.0 | 42.0 | 26.3 | 14.7 | 49.9 |
| $45.7 - 50.8$ | 1.17 | 0.419 | 10653 | 121.9 | 63.5 | 44.3 | 13.8 | 89.2 |
| $50.8 - 55.9$ | 1.17 | 0.451 | 11610 | 126.4 | 70.0 | 48.9 | 7.5 | 101.1 |
| $55.9 - 61.0$ | 1.16 | 0.473 | 9379 | 111.4 | 63.5 | 42.0 | 6.4 | 89.5 |

Table 7. Measured bulk density, water content, electrical conductivity, and ion concentration profiles for conditions of case $#3$

| Depth | EC | Total Salts | Calcium | Magnesium | Sodium | Chloride |
|---------------------------------------|--------------|----------------|------------------|-----------|--------|----------|
| (cm) | $(\mu$ mhos) | (me/1) | (me/1) | (me/1) | (me/1) | (me/1) |
| | | | 324 hours (a) | | | |
| 30 $0 -$ | 3281 | 42.9 | 18.5 | 11.3 | 13.1 | 16.2 |
| 60 $30 -$ | 6952 | 72.9 | 32.4 | 41.1 | 8.4 | 57.6 |
| 60 90 $\overline{}$ | 2941 | 41.4 | 12.8 | 22.4 | 6.3 | 3.7 |
| 90 120 $\overline{}$ | 3178 | 46.0 | 16.3 | 23.9 | 5.9 | 0.4 |
| | | | 329 hours (b) | | | |
| $0 -$ 30 | 4571 | 59.1 | 12.3 | 6.9 | 40.0 | 14.1 |
| $30 -$ 60 | 6996 | 70.6 | 22.5 | 41.7 | 6.4 | 51.9 |
| 60 90 $\overline{}$ | 4562 | 48.1 | 14.9 | 26.5 | 6.6 | 13.8 |
| 90 120 $\overline{}$ | 3670 | 47.9 | 14.3 | 27.2 | 6.4 | 2.5 |
| | | | (c) 627 hours | | | |
| $0 -$ 30 | 6468 | 73.0 | 21.1 | 6.7 | 45.2 | 40.2 |
| 60 30 $\frac{1}{2}$ | 8981 | 88.6 | 46.7 | 33.2 | 8.7 | 78.1 |
| 60 90 $\overline{}$ | 7260 | 73.6 | 34.1 | 34.8 | 4.8 | 53.3 |
| 90 120 $\qquad \qquad -$ | 4462 | 52.8 | 30.6 | 17.7 | 4.6 | 16.4 |

Table 8. Chemical analysis of saturation extract for field experiment

| | | | Time | | | | |
|-------|----------|--------|---------|---------|------------|--|--|
| Depth | 24 hrs | 40 hrs | 324 hrs | 339 hrs | 627 hrs | | |
| (cm) | | | | | | | |
| 30 | 0.20 | 0.32 | 0.24 | 0.32 | 0.28 | | |
| 45 | 0.24 | 0.32 | 0.28 | 0.33 | 0.29 | | |
| 75 | 0.25 | 0.32 | 0.27 | 0.33 | 0.28 | | |
| 105 | 0.29 | 0.33 | 0.31 | 0.34 | 0.31 | | |
| 135 | 0.38 | 0.40 | 0.39 | 0.40 | 0.39 | | |
| 165 | 0.41 | 0.42 | 0.42 | 0.42 | 0.41 | | |
| | | | | | | | |

Table 9. Water content (θ) profiles for field experiment

Table 10. Electrical conductivity (μ mhos) profiles at the field water **content**

| | Time | | | | | | | | | |
|-------|----------|--------|--------|-----------|-----------|-----------|--|--|--|--|
| Depth | 24 hrs | 48 hrs | 76 hrs | 324 hrs | 327 hrs | 627 hrs | | | | |
| (cm) | | | | | | | | | | |
| 15 | | 4082 | 3740 | 6915 | 8747 | 8883 | | | | |
| 45 | 3369 | 33657 | 32422 | 15070 | 14745 | 14350 | | | | |
| 75 | 4068 | 4252 | 13877 | 7032 | 7175 | 9285 | | | | |
| 105 | 3111 | 6686 | 6382 | 3930 | 4410 | 4756 | | | | |
| 135 | 3426 | 3279 | 3012 | 3372 | 3594 | 4373 | | | | |
| 165 | 2671 | 2820 | 2765 | 2654 | 2700 | 2622 | | | | |

Appendix B. FORTRAN Program

PROGRAM-SOIL WATER . SALT . FLOW WITH PLANT UPTAKE $\mathbf c$ PROGRAM OF SEP. 25.1971 \mathbf{c} HWET IS PRESSURE OF HIGHEST POSSIBLE WATER CONTENT C V IS BOUNDARY CONDITIONS AT TOP AND TIMES CONDITIONS APPLY ϵ DETT IS TIME INCREMENT TO START WITH AND LOWEST TO USE C CONG IS SMALLEST WATERCONTENT CHANGE ALLOWED EACH COMPUTATION c GRAVY IS GRAVITY COMPONENT USUALLY THE SAME AS DELX ϵ DELW IS WATER CONTENT DIFFERENCE CORRESPONDING TO TABLE INCREMENTS ϵ T IS WATER CONTENT TARLE HAS EQUAL SPACED INCREMENTS ϵ TIME IS CUMULATIVE TIME AT START OF COMPUTATION ϵ TT IS 1.0 FOR LAASONEN AND 0.5 FOR CRANK NICHOLSON \mathfrak{c} CUMT IS TIME AT END OF COMPUTATION ϵ TAA=1. FOR ZERO FLUX AT BOTTOM. TAA=0 FOR HIKKI CONSTANT Λ ϵ FROM GIII OR HIII=GIII C CTM IS LOWEST VALUE OF DELT PERMITTED--IF AS LOW STOPS \mathbf{r} HORY IS PRESSURE OF LOWEST POSSIBLE WATER CONTENT \mathbf{c} PP IS PRESSURE TABLE (WETTING) STARTING WITH LOWEST PRESSURE ϵ D IS CONDUCTIVITY TABLE STARTING WITH LOWEST WATER CONTENT VALUE
QQ SAME AS ABOVE EXCEPT STARTS FROM WETTING \mathbf{c} \mathbf{c} C IS WATER CAPACITY AS A FUNCTION OF DEPTH BEGINNING AT TOP $\mathbf c$ DELX IS DEPTH INCREMENT \mathfrak{c} W IS WATER CONTENT AS A FUNCTION OF DEPTH BEGINNING AT TOP C H IS WATER PRESSURE AS A FUNCTION OF DEPTH BEGINNING AT TOP ϵ WATL IS LOWEST POSSIBLE WATER CONTENT \mathbb{C} WATH IS HIGHEST POSSIBLE WATER CONTENT ϵ CB IS A CONSTANT TO MULTIPLY D ARRAY BY--USUALLY 1.0 ϵ K IS NO. OF DELX INCREMENTS.MM NO. OF TIMES HOW PRINTED.KIT NO.OF 12 Δ Γ $\overline{\mathbf{A}}$ 13 - START HERE FOR A NEW PROGRAM C Δ 14 ϵ MI IS TO PRINT H.W ARRAYS EACH ITER..IER NO. OF V ELEMENTS HROOT IS THE ACTUAL ROOT WATER POTENTIAL \mathbf{c} BR REPRESENTS PLANT UPTAKE ADDITIONS \mathbf{c} HLOW IS THE MINIMUM ROOT POTENTIAL ALLOWED \mathbf{c} HHI IS THE MAXIMIM ROOT POTENTIAL ALLOWED C ET IS THE POTENTIAL EVAPOTRANSPIRATION. AL WAYS NEG ATIVE ϵ WFDD IS THE WATER FLOW RATE AT THE SURFACE \mathcal{C} ETPL IS THE POTENTIAL TRANSPIRATION +ALWAYS NEGATIVE C \mathbf{c} SUMS-SALT CONCENTRATION GOING OUT TET IS THE BOUNDARY POTENTIAL ET. ALWAYS NEGATIVE.LTV ARRAY C DD.H.G.Y.W.RDF.A.SE.SS.SD ARRAYS ARE OF SAME DIMENSION AT LEAST =KK c P.D.T.ARE OF EQUAL DIMENSIONS.ARE EQUAL TO 60 AT MOST \overline{c} SF. TET.V ARRAYS ARE OF SAME DIMENSION AT LEAST =IER $\mathbf c$ CS=CALCIUM . MS=MAGNESIUM . SN=SODIUM . CL=CHLORIDE . SU= SULP HATE . HC= BICARB ON ATE \mathbf{c} CE.ME.EN.ARE EXCHANGBLE CALCIUM.MAGNESIUM.SODIUM ϵ CAL.CAS.ARE CALCITE. \mathfrak{g} CAL . CAS.ARE CALCITE AND GYPSUM. CSO.MGSO.ARE ION PAIRS $\mathsf{c}\,$ CSG.MSO.SNO.CLO.HCO.SUO.CSX.MGSX.ARE OLD CONCENTRATIOS ϵ AM.AC.SAM.SAC.ARE TABLES OF IONIC-STRENGTH ACTIVITY COEFFICIENT OF C ϵ CALCIUM AND SODIUM IONS TCA . TMG . ARE TOTALCONCER AT IONS (CATIONS . ION - P AIR S) \mathfrak{c} DIMENSION CF(35)+MF(35)+SNF(35)+SUF(35)+CSOF(35)+MGSF(35)+CLF(35) DIMENSION HCF (35) DIMENSION DD(25) . H(25) . G(25) . Y(25) . W(25) . RDF(25) . A(25) . SE(25) DIMENSION SS(25) . SD(25) . C(25) . B(25) . E(25) . F(25) DIMENSION SFI651.TETI651.VI651 DIMENSION P(50) . D(50) . T (50) DIMENSION CS(27)+MS(27)+SN(27)+SU(27)+CL(27)+HC(27)+CE(26)+ME(26) DIMENSION EN(26)+CAS(27)+CSO(27)+MGSX(27)+TMG(27)

```
DIMENSION CSG(27)+MS0(27)+SN0(27)+CL0(27)+HC0(27)+SU0(27)+CSX(27)
     DIMENSION AM(27)+AC(27)+SAM(26)+SAC(26)
     DIMENSION TCA(26)
     DIMENSION CAL (27) .MGSOL27)
     REAL MF.MGSF
     REAL MGSO.MGSX.MSP
     REAL MSO
     REAL MS.ME.MEA.MSA
     WRITE (6+8765)
.............<sup>1</sup>
    WRITE (6 . 8766)
8766 FORMATILH . 34 X. * CROP ALFALFA.. ROOT DEPTH IS (2) FEET. *)
     WRITE (6,9999)
9999 FORMAT(1H +25X+ *DATA ARE FROM 3/8/1971 TO 11/9/1971 CROP ALFALFA+ E
    1 - 0 - 10ET<sup>*</sup>)
     WRITE (6+8767)
.............!
                                                                           \Delta\mathbf{q}PEAD 163. ML
    LMM = 0\Delta11\mathtt{A}10
    LMM=LMM+1
     READ 163. K.MM. IER.NB.ND
                                                                           \Delta18
     KK = K + 1READ 165, (DD(I), I=1, KK)
     READ(5,165) (TET(I),T=1,1ER)
     READ(5,165) (PDF(I),I=1,KK)
     KC = 1ET = TET(1)\mathtt{A}19
     LL =MM
     READ(5, 165) (P(I), I=1, ND)
     READ(5, 165)(D(I), I=1, ND)
                                                                           \mathbf{A}25READ 165. (W(I).I=1.KK)
                                                                           \Delta26READ 165, (VII), I=1, IER)
     READ 165, DELX, DETT, CRAVY, CONG, DELW, TIME
     READ 165, TT, CUMT, TAA, HLOW, HHI, RRES
     READ 165. HDRY. HWET.WATL. WATH.CB
     BOUNDRY CONDITIONS FOR SALT FLOW
     READ 165, (SF(I), I=1, IER)
     READIS, 165) (CF(I), I=1, IER)
     READIS, 1651 (MF (I), I=1, IER)
     READ (5, 165) (SNF (I), I=1, IER)
     READ(5, 165)(SUF(I), I=1, IER)
     READ(5+165)(CLF(I)+I=1+IER)
     READ(5, 165) (HCF(I), I=1, IER)
     READ(5, 165)(CSOF(I), I=1, IER)
     READ(5+165)(MGSF(I)+I=1+IER)
     WRITE(6,166)(CF(I), I=1, IER)
     WRITE(6+166)(MF(I)+I=1+IER)
     WRITE(6+166)(SNF(I)+I=1+IER)
     WRITE(6.166)(SUF(I).I=1.TER)
     WRITE(6+166)(CLF(I)+I=1+IER)
     WRITE(6,166)(HCF(I),I=1,IER)
     WRITE(6+166)(CSOF(I)+I=1+IER)
     WRITE(6+166)(MGSF(I)+I=1+IER)
     READIS, 165) AM
     READ(5, 165)AC
     READ(5, 165)SAM
     READIS, 1651SAC
     WRITE(6,166)(AM(I), AC(I), I=1, 27)
     WRITE(6+166)(SAM(I)+SAC(I)+I=1+26)
```
 $\mathbf 1$

 \mathbf{C}

```
INTIAL CONDITIONS FOR SALT FLOW
\mathsf{C}READIS, 1651(CS(I), I=1, KK)
      READIS. 1651 (MS(I).I=1.KK)
      READIS, 165) (SN(I), I=1, KK)
      READIS, 1651(SU(I), I=1, KK)
      READIS, 165) (CLIII, I=1, KK)
      READ (5+165) (HC(I)+I=1+KK)
      READ(5,165)(CAL(I), I=1, KK)
      READ(5, 165)(CAS(I), I=1, KK)
      WRITE(6,166)(CS(I), I=1, KK)
      WRITE(6,166)(MS(I), I=1, KK)
      WRITE(6+166)(SN(I)+I=1+KK)
      WRITE(6,166)(SU(I),I=1,KK)
      WRITE(6,166)(CL(1), I=1, KK)
      WRITE(6+166)(HC(I)+I=1+KK)
      WRITE(6,166)(CAL(I),I=1,KK)
      WRITE(6+166)(CAS(I)+I=1+KK)
      WRITE (6, 169)
      WRITE (6,163) K.MM. IER. NB.ND
      SMAXIWATH * 4 OD.
      P(1) = P(1) * 1 * 0E * 03\Delta38
      1(1) = 0.0STIM = 0.0DO 900 I=2.ND
      T(I)=DELW+T(I-1)
900 P(I)=P(I)+1.0E+03
                                                                     -------
C - -00 600 I=2.KK
      CSA=CS(I)
      MSAIMS(I)
      SNA=SNII)
      SUA=SU(I)
      CLA=CL(I)
      HCA=HC(I)
      SEA=CS(I)+MS(I)+SN(I)
      CALL EGEXCHICSA.MSA.SNA.CLA.SUA.HCA.CEA.MEA.ENA.CSP.MSP.SEA.AM.AC.
     QSAM.SACI
      CSITIECSA
      MSITI-MSA
      SNITTSNA
      SUITI-SUA
      CL (I) =CLA
      HC(I)=HCA
      CEITI=CEA
      MEITIEMEA
      ENCIJ=ENA
      CSO(I)=CSP
      MGSOIII=MSP
      SEITTESEA
      WRITE(6+166)CS(I)+MS(I)+SN(I)+CL(I)+SU(I)+HC(I)+CE(I)+ME(I)+EN(I)+
     GCSO(I),MGSO(I)
 600 CONTINUE
      CS(2)=2590.9588CL (2) = 2590-9588
  825 WRITE (6.911)
      DO 910 I=1.KK
  910 WRITE(6+166)(CS(I)+MS(I)+SN(I)+SU(I)+CL(I)+HC(I)+CE(I)+ME(I)+EN(I)
     1.CAS(I).CSO(I))
      WRITE (6+166)MGSO
C - - - -SE(11=SE(1)
```



```
1/(DD(T+1)-DD(T))WERD=(B(I+1)+((H(I+1)-H(I+2))+TT+(G(I+1)-G(I+2))+TM+DD(I+2)-DD(I+1
   JJ) *DELTI/(DD(I+2)-DD(I+1))
254 IF(ABS(WFRU).LT.0.0001.AND.ABS(WFRD).LT.0.0001)60 TO 920
    CLI = CLO(1)CL2=CL0(I+1)CL3=CL0(I+2)CL4=W1I+1CL5=Y(I+1)CALL SALT (CL1+CL2+CL3+CL4+CL5+CL6+WFRU+WFRD+DELX+I+EOR)
    CL(I+1)=CL6C1.15500 (T)
    CL2 = SUO(I+1)CL3 = SUO(I+2)CALL SALT ICL 1+CL2+CL 3+CL 4+CL 5+CL 6+WFRU+WFRD+DELX+I+EOR)
    SU(I+1)=CL6CL 1 = HCO(1)CL2=HCO(I+1)CL3=HC0(I+2)CALL SALT (CL1+CL2+CL3+CL4+CL5+CL6+WFRU+WFRD+DELX+I+EOR)
    HC(I+1)=CL6
    CL1 = C56(1)CL2 = C56(1+1)CL3 = C5611 * 21CALL SALT (CL1+CL2+CL3+CL4+CL5+CL6+WFRU+WFRD+DELX+I+EOR)
    CS(I+1)=CL6CL1=MSO(1)CL2=MSO(I+1)CL3=MS0(I+2)
    CALL SALT (CL1+CL2+CL3+CL4+CL5+CL6+WFRU+WFRD+DELX+1+EOR)
    MS(1+1)=CL6CL1 = SNO(1)CL2 = SNO(I+1)CL3 = SNO(I*2)CALL SALT (CL1+CL2+CL3+CL4+CL5+CL6+WFRU+WFRD+DELX+I+EOR)
    SN(I+1)=CL6CL1=CSX(I)
    CL2=CSX(I+1)CL3=CSX(I+2)CALL SALT (CL1+CL2+CL3+CL4+CL5+CL6+WFRU+WFRD+DELX+I+EOR)
    CSO(1+1) = CL6
    CL1 = MGSX(1)CL2=MGSX(I+1)CL3=MGSX(I+2)
    CALL SALT (CL1+CL2+CL3+CL4+CL5+CL6+WFRU+WFRD+DELX+I+EOR)
    MGSO(I+1)=CL6SE (I+1)=CS(I+1)+MS(I+1)+SN(I+1)+CS0(I+1)+MGS0(I+1);
    TCA (I+1) = C S (I+1) + C S O (I+1)TMG(I+1)=MS(I+1)+MGS0(I+1)
    LS = I + 1IF(STIM.LT.1.0160 TO 920
475 WATEW(I+1)CSA=CS(I+1)MSA=MS(I+1)
    SNA = SN(I + 1)SUB=SUTE+11CLA = CL (I + 1)
    HCA = HC(T + 1)CEA = CE(I+1)MEATME (I+1)ENA=EN(I+1)
```


TW=1000.0*DET


```
SURROUTINE EXCHIA.F.S.G.H.HCO 3. ET.CT.SAT. CASO.AGS O. XX T. CAL. PWI.SEA
     M.AM.AC.SAM.SACI
      DIMENSION AM(27)+AC(27)+SAM(26)+SAC(26)
       DOUBLE PRECISION ZZZ
       MH = 1IG=1PW1=PW1+100-0/1-16
       8=100000.U/PW1
       B1 = P W1A=A/(1000.0*2.0)
       F = F / (1000 - 0 * 2 * 0)G = G / (1000 - 0 * 2 - 0)5 = 5/1000 - 0H = H / 1000 - 0HCO3=HCO3/1000.
       4650=4650/(1000-*2-)
       CASO=CASO/11000.02.1
\mathsf{c}DA IS KICA-NAIEXCHANGE COEFFCIENT
       DA=4.5377
       D = 0.25 * 5.0/SFA
       U=50RT(2.0*(A+F+G)+0.5*(S+H+HC03))
       IF(U**2.LT.0.003)60 TO 100
       CALL ACOF(AM, AC, SAM, SAC, U, ADI, AMO)
       ASA=ADI . AMO . . 2.0
       GO TO 101
  INU ASA=EXP(-7.0242*U/(1.0*U))
  101 IF(CAL) 1000+602+603
 602TK = 1AAA=452.
  FOR ZE = AAA/(B1**1.68)
       GO TO 24
  603 IK=2
       ZE = I - 1.68 * ALOG(B1) - 4.46) * 2.3
       ZE=EXPIZE)/ASA
       A1 = A24IF (XXT) 4.4, 26
      U=SGRT (2.0*(A+F+G)+0.5*(S+H+HCO3))
 4
       IF(U**2.LT.0.003)60 TO 102
      CALL ACOF (AM+ AC+SAM+SAC+U+ADI+AMO)
       AA=ADI ** 2
       GO TO 103
  IN 2 AA =EXP (-9.366 *U/(1.0*U))
  IN3 IF (2.4E-5-A*6*AA) 26.18.18
  25 -x=0-0U=SQRT(2.0*(A+F+G)+0.5*(S+H+HCO3))<br>IF(U**2.LT.0.003)60 TO 104
       CALL ACOF(AM+AC+SAM+SAC+U+ADI+AM0)
       EX=1.0/(ADI ** 2)
       GO TO 105
  IN4 EX=EXP((9.366*U)/(1.0*U))
  105 B8 = A + GCC = A * G - (2 - 4E - 5) * EXR=SORT(BB+BB-4.D*CC)
       x = (-B B + R) / 2.0CAS1 = 4 - 897E - 3 - CAS0DEL=R*XXT-CAS1
       IF (DEL-X) 27.2 R. 28
  27 X=XXT*Bx \times 1 = 0.0CAS1 = 0.0A = A + X
```

```
G = G * XU=SORT(2.0*(A+F+G)+0.5*(S+H+HC03))
      IF(U**2.LT.0.003)60 TO 106
      CALL ACOF (AM. AC. SAM. SAC.U.ADI.AMO)
      AA=ADI .. 2
      GO TO 7
 106 AA=EXP (-9.366*U/(1.*U))
     BB = -14 - 9E - 3 + AA + A + AA + 6\overline{7}CC = AA * A * G - 4 * 9E - 3 * C A S0XXXX = BB * BB - 4 - 0 * A + CCIF (XXXX)35.35.36
 35x1=0.0GO TO 37
     X1=(-BB-SGRI(XXXX))/(2.0*AA)
 TE
     CASO=CASO+X1
 77A = A - X]
      G = G - X1GO TO 44
 19IF (G) 1.1.6IF (A) 1. 1.7F\mathbf{1}IF (CASO)44.44.7
 29A = A \cdot X6 = 6 * XXXT = XXT - X \times BCASO=CASO+CAS1
      XXT=XXT-CAS1/B
 44 A2=A
     I<sub>I</sub>=1I H = 1MX = 1IF(5)80.181.80
 131 IF (SAT) 80 .515 .80
     IJ=2BD404 IF (SAT-ET1402+403+403
 4N2 Z=SAT/10.
      21 - 7GO TO 5
 403 Z=ET/10.
     Z1=ZU=SQRT(2.0+(A+F+G)+0.5+(S+H+HCO3))<br>IF(U++2.LT.0.003)60 T0 107
 \mathsf{S}CALL ACOFIAM.AC.SAM.SAC.U.ADI.AMOI
      EX=ADI/AMO**2
      GO TO 108
 INT EX=EXP((-2.341*U)/(1.0*U))
 108 AA = -4.0*DA*DA*B*B\texttt{BB=4-0*B*1EX+2-0*DA*DA*E7*B*DA*DA*S1}CC = 4.7* EX* (A*SAT*B) - 4.0*DA*DA*BET* (B*ET*2.0*5)-DA*DA*S*S\texttt{DD=SATEE} \times \texttt{14.0*A+SAT*B1+2.0*DA*DA*E7*S*12.0*B*E7*S1}EE=SAT*SAT*A*EX-DA*DA*S*S*ET*ET
     ZZ=-IIIIAA+Z+BBI+Z+CCI+Z+DDI+Z+EEI
 81
      ZZZ=ttt4.0+AA+Z+3.0+BB)+Z+2.0+CC)+Z+DD1
      77 - 7717772IF (27-0.0)302,303,302
303 IF(2-0-0)302-515-302
 302 222=22/2
      Z = Z + ZZIF(IH.GE.5)GO TO 83
      IH = IH + 1IF(DABS(ZZZ)-0.001)83.83.81
83 IF1Z.GE.0.0160 TO 305
```

```
IF (SAT.LT.ABS(Z+2.0))Z=SAT/2.0
 315 A=A+B+7
     IF (A) 510+ 510+ 512
 552 SAT=SAT-2.+2
 551 ET=ET+Z
 550 S=S+2.+B+Z
 510 A = A - B * ZZ = -Z1IF (MX.GE.5) GO TO 512
     MX = MX + 1GO TO 81
 51.2 5=5-2.0 *B * Z
     IF (MX.GE.5) GO TO 513
     MX=MX+1IF (5) 550, 550, 513
 51 3 ET=ET-7IF (MX. GE. 5) GO TO 514
     MX=MX+1IF (ET) 551 .551 .514
 514 SAT=SAT+2-0+7
     IF (MX.GE.5)GO TO 515
     MX = MX + 1IF (SAT) 55 2, 55 2, 51 5
 51 5 A 3 = A
     BB=A+B+(CT+D+ET)+D+F
     AA = B * (1 - 0 - 0)CC = (A*CT - D*F*ET)R=SORT(BB+BB-4.0+AA+CC)
     Y = (-BB + R) / (2 - 0 * AA)A = A \cdot B \cdot YFIF - B*YET = FT - YCT = CT + YIF (G) 790.790.791
 791 IF (F1790, 790, 792)
 792 IF(U**2.LT.0.003)60 TO 109
     CALL ACOFIAM. AC.SAM.SAC.U.ADI.AMOI
     AA = ADI * * 2GO TO 110
 109 AA=EXP(-9.366+U/(1.+U))
 11 U BB = - 15 - 9E - 3 + A \Delta * F + A \Delta * GCC = AA * F * G - 5 * 9E - 3 * A G S0XXXX=BB*BB-4.0*AA+CC
     IF (XXXX) 793.793.794
 79.3 \times 1 = 0.0GO TO 795
 794 X1=(-BB-SQRT(XXXX))/(2.0*AA)
795 AGSO=AGSO+X1
     F = F - X1
     G = G - X179U CONTINUE
     60 TO (600+601) .IK
 601 \, 14129.0BB=4.0*HC03+A
     CC=HC03++2+4.0+A+HC03
     DD = A * HC 03 * * 2 - ZEIF (HCO3-A)61.61.62
 61
    Z = -HCO3/460 TO 650
 62 \t 7=-A/265U 21=2
```
63 ZZ=-IIIAA+Z+BBI+Z+CCJ+Z+DDI 222=(13.0*AA*Z*2.0*BB)*Z*CC) IF(22-0.0)300.301.300 301 IF (222-0.0) 300.600.300 300 22=22/222 $277 = 2777$ $7 - 7 + 77$ IF(IT.6E.5)60 TO 64 $IT = IT + 1$ IF (DABS (222) - - 001164, 64.63 $A = A + 7$ 64 $HCO3=HC03*2*2$ IF (HC03)752,752,651 752 HCO3=HCO3-2.02 $A = A - 7$ $2 = -21$ GO TO 63 651 IF (A) 752.752.753 753 CAL=CAL-Z SNU ZX=A+HC03++2 IF (2x-7E) 606, 605, 605 505 $IK = 2$ 676 DEL=A-A1 IF(16.6E.5)60 TO 8 $IG = I6 + 1$ IF(DEL+1.0E-5124.48.48 IF(DEL-1.DE-5149.49.24 \mathbf{a} $DEL = A - A2$ 49 IF(DEL+1.0E-5)24,50,50 IF (DEL-1.0E-5151+51+24) 50 51 $DEL = A - A$ IF ($DEL + 1.0E - 5124.52.52$ 52 IF(DEL-1.DE-518+8+24) $\overline{8}$ $AP = A$ 1000 CONTINUE 67 CONTINUE A=A * 1000.0*2.0 $F = F * 1000 - 0 * 2 - 0$ $S = S * 1000 - 0$ $6:6*1000.0*2.0$ H=H+1000.0 HCO 3=HCO3 *1000. $CASO = CASO * 1000 - D * 2 - D$ A650=A650*1000+*2-0 SEA=A+F+S+CASO+AGSO RETURN **FND**

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SUBROUTINE EGEXCHICA+AMG+SOS+CL+SO+HCO3+E5+C5+SA5+CASO+AGSO+SEA+AM Y.AC.SAM.SACI DIMENSION AM(27)+AC(27)+SAM(26)+SAC(26) $CA = CA / (1000 - 0 * 2 - 0)$ AMG=AMG/(1000-0*2.0) 505=505/1000. $50:50/11000.0*2.01$ $CL = CL/1000$. HCO 3=HCO3/1000-0 $EC = 0 - 11E - 03$ DA IS KINA-CAI OR 1.0/KICA-NAI EXCHANGE COEFFCIENT

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DA=0.3115 $D = D - 25 + 5 - D/5F$ A $CASO = D - D$ U=SORT(2.0+(CA+AMG+S0)+0.5+(S0S+HC03+CL)) $AGSO=0.0$ IF (U**2.LT.0.003)60 TO 42 CALL ACOF (AM. AC.SAM.SAC.U.ADI.AMO) $ACI22ADI+2$ GO TO 150 42 ACT2=EXP(-9.366*U/(1.0*U)) 15U IF (50)1000+713+712 712 AA=ACT2 * ACT2 BB=ACT2+(10-8E-3+(ACT2+(AMG+CA-50))) $CC = 28.91E - 6 + (ACT2 * (AMG * 4.9E - 3 * (CA * 5.9E - 3) - (50 * 10.8E - 3))$ $DD = -50 \cdot 28 - 91F - 6$ $8002550/2$ $85 \text{ u } 21 = 2$ 863 27 = - (((AA*Z*BB)*Z*CC)*Z*DD) $777 - 113 - 0 * AA * 7 * 2 - 0 * B B * 2 * CC$ $22 = 22/222$ 77722717 $7 - 7 + 77$ IF (ABS(222)-.001)840, R40, 863 **R4 U SOT=SO** $50 - 7$ IF (50) 710 . 710 . 711 71 0 SO=SOT $Z = Z1$ GO TO 863 711 CASX=SO*CA*ACT2/14.9F-3*ACT2*SO1 $C X = C A - C A S X$ AGSX=SO+AMG+ACT2/15.9E-3+ACT2+50) $AMX = AMG - AGSX$ UU=SORT(2.0*(CX+AMX+S0)+0.5*(SOS+HC03+CL)) TE CARS CULL/U-1-1-1-0E-4140.40.41 41 U=UU $50 = 50$ GO TO 42 40 CASO=CASX **AGSO=AGSX** $C = CX$ **AMGIAMX** 713 ACT1=SORT(ACT2) ACTM=SORT(ACT1) ACTM=SORT(ACTM) $CA = CA + 2$ AMG=AMG+2. C--CA.MG.ARE IN EQUIVALENT/LITER ES=EC/ILACTM=SOS/LDA=SQRTLACT 1+CA))) +1.+ LD+ ACT1+AMG/LACT1 +CA))) SAS=ACTM*SOS*E5/ISORTIACT1*CA1*DA1 $C5 = EC-E5-SAS$ $E5 = E5/2$. $C5 = C5/2$. $CA = CA * 1000 - 0$ AM6=AM6+1000.0 C--CA.MG.ARE IN ME/LITER 505=505+1000.0 $CL = CL + 1000 - 0$ 50=50*1000.0*2.0 HCO 3=HCO3 *1000.0 CASO=CASO*1000.0*2.0

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SUBROUTINE SALT(A+B+C+D+E+F+WFRU+WFRD+DELX+J+EOR)
      SE AND SS ARE NEW AND OLD CONCENTRATIONS
\mathsf{r}DIMENSION SS(3) .W(3), SE(3), Y(3)
      \mathbf{I} = \mathbf{1}SS(1)=A55(1+1)=8SS(1+2)=CW(I+1)=0\gamma ( \texttt{I} + \texttt{l} ) \texttt{E}IF(J.E0.1160 TO 201
      IF ( WFRU. GE. D. D. AND. WFRD. GE. 0.0160 TO 205
      IF (WFRU.LE.U.D.AND.WFRD.LE.D.D)GO TO 209
      IF (WFRU.GE.D.D.AND.WFRD.LE.D.D)GO TO 208
      IF(WFRU.LE.D.O.AND.WFRD.GE.D.D)GO TO 210
  205 SE (1+1)=(SS(1+1)+Y(1+1)+(SS(1)+WFRU-SS(1+1)+WFRD)/DELX)/W(1+1)
      GO TO 200
  201 IF(E0R-0.0)203,204,202
     IF (WFRD.GT.D.D) GO TO 206
 203
      SE ( 2) = ( SS( 2) + Y ( 2) - ( SS( 3) + WFRD) / DEL X ) /W ( 2)
 207GO TO 200
  206 SE121=155121*Y121-155121*WFRD1/DELX1/W121
      GO TO 200
  204 IF (WFRD.LT.D.D) GO TO 207
      GO TO 206
  202 IF (WFRD.6T.D.D) GO TO 205
  208 SE (I+1)=(SS(I+1)+Y(I+1)+(SS(I)+WFRU-SS(I+2)+WFRD)/DELX)/W(I+1)
      GO TO 200
  209 SE (I+1)=(SS(I+1)+Y(I+1)+(SS(I+1)+WFRU-SS(I+2)+WFRD)/DELX)/W(I+1)
      GO TO 200
  210 SE (1+1)=(SS(1+1)*Y(1+1)+(SS(1+1)*WFRU-SS(1+1)*WFR0)/DELX)/W(1+1)
  200 IF (SE (I+1).LT.0.0) SE (I+1)=SS(I+1)
      F = SE(I+1)RETURN
      END
      SUBROUTINE ACOFIAM, AC.SAM.SAC.U.ADI.AMOI
      DIMENSION AM(27), AC(27), SAM(26), SAC(26)
      U = U * * 2IF(U.LE.15.0)60 TO 1
      ADI = 587 - DGO TO 2
      DO 751 I=1.27
  \mathbf{1}IF (U.GT. AM (I) 160 TO 751
      AG=U-AM(I-1)ADI=AC(I-1)+((AC(I)-AC(I-1))+AG)/(AM(I)-AM(I-1))
      GO TO 2
  751 CONTINUE
```
WRITE(6+100)CA+AMG+SOS+CL+SO+HCO3+E5+C5+SA5+AGSO+CASO+SEA

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\overline{z}IF (U.LE.5.0)60 TO 3
     AMO = 1.295
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AGS0=AGS0+1000.U+2.0 SEA=CA+AMG+SOS+CASO+AGSO

100 FORMAT (12E10-4) 1000 RETURN END

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\overline{\mathfrak{z}}AG=U-SAM(I-1)<br>AMO=SAC(I-1)+((SAC(I)-SAC(I-1))+AG)/(SAM(I)-SAM(I-1)3<br>GO TO 4<br>CONTINUE<br>RETURN<br>RETURN
6C4
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VITA

Satish Chander Gupta Candidate for the degree of Doctor of Philosophy

Dissertation: Model for predicting simultaneous distribution of salt and water in soils

Major Field: Soils and Biometeorology

Biographical Information:

Personal data: Born at Poonch, Jammu India, July 29, 1945; son of Sita Ram and Laj Wanti Gupta.

- Education: Received the Bachelor of Science degree from Punjab Agricultural University, Ludhiana, majoring in Soil Science, in 1966; received the Master of Science degree from Punjab Agricultural University, Hissar, majoring in Soil Science, in 1968; completed requirements for the Doctor of Philosophy majoring in Soil Physics, at Utah State University, Logan, Utah in 1972.
- Professional Activities: Selected for junior research fellowship 1966-1968 by the Indian Council of Agricultural Research; paper presented: "Influence of water content on electrical conductivity of the soil." Annual meeting of the Western Soil Science Society of America, 1971.
- Affiliations: American Society of Agronomy; Soil Science Society of America.