

1  
2  
3  
4  
5  
6  
7  
8  
9  
10

**Season- and herbivore-dependent competition and facilitation  
in a semi-arid savanna**

Kari E. Veblen

Dept of Plant Sciences and Graduate Group in Ecology

University of California

Davis, CA 95616

keveblen@ucdavis.edu

**11 Abstract**

12 Empirical and theoretical evidence suggests that facilitation between plants, when it occurs, is  
13 more likely during periods of abiotic stress, while competition predominates under more  
14 moderate conditions. Therefore, one might expect the relative importance of competition vs.  
15 facilitation to vary seasonally in ecosystems characterized by pronounced dry (abiotically  
16 stressful) and wet (benign) seasons. Herbivory also varies seasonally and can affect the net  
17 outcome of plant-plant interactions, but the interactive effects of seasonality and herbivory on the  
18 competition-facilitation balance are not known. I experimentally manipulated neighboring plants  
19 and herbivory during wet and dry periods for two species of grass: *Cynodon plectostachyus* and  
20 *Pennisetum stramineum*, in the semi-arid Laikipia District of Kenya. These experiments  
21 indicate that *Pennisetum* was competitively dominant during the wet season and that it responded  
22 negatively to grazing, especially during the dry season. *Cynodon* showed more complex season-  
23 and herbivore-dependent responses. *Cynodon* experienced facilitation that was simultaneously  
24 dependent on presence of herbivores and on dry season. During the wet season *Cynodon*  
25 experienced net competition. These results illustrate how herbivory and seasonality can interact  
26 in complex ways to shift species-species competition-facilitation balance. Additionally, because  
27 *Cynodon* and *Pennisetum* are key players in a local successional process, these results indicate  
28 that herbivory can affect the direction and pace of succession.

29

**30 Key words**

31 *Acacia drepanolobium* savanna; associational resistance; *Cynodon plectostachyus*; grasslands;  
32 facilitation; indirect effects; herbivory; *Pennisetum stramineum*; succession

33

## 34 **Introduction**

35           Plant communities are shaped by complex interactions among multiple biotic and abiotic  
36 factors. Although competition has been emphasized as the primary biotic driver of plant  
37 interactions (Connell 1983, Schoener 1983, Goldberg and Barton 1992), a more recent profusion  
38 of evidence indicates that net plant-plant interactions often are the result of both competitive and  
39 facilitative interactions (Callaway 1995, Holmgren et al. 1997, Kikvidze et al. 2006, Graff et al.  
40 2007). Theoretical and empirical evidence suggests that facilitative (positive) relationships are  
41 more common in stressful or disturbed environments where the facilitator ameliorates abiotic  
42 stress (Pugnaire and Luque 2001, Callaway et al. 2002) or consumer stress (Holmes and Jepson-  
43 Innes 1989, Bertness and Callaway 1994, Gómez et al. 2001). Not surprisingly, a preponderance  
44 of examples of facilitation in plants, both direct (Callaway 1992, Maestre et al. 2003) and  
45 indirect (McNaughton 1978, Holmes and Jepson-Innes 1989, Gómez et al. 2001, Rebollo et al.  
46 2002) are found in arid and semi-arid systems that undergo extremes in both abiotic stress (e.g.,  
47 drought) and consumer stress (e.g., herbivory). Surprisingly few studies, however, have  
48 examined interactions between these two stresses.

49           Based on the abiotic stress hypothesis, the net direction of plant-plant interactions in  
50 water-limited environments would be expected to vary seasonally, according to water  
51 availability. In fact, several studies have found that the competition-facilitation balance varies  
52 according to annual water availability (Greenlee and Callaway 1996, Pugnaire and Lazaro 2000,  
53 Tielborger and Kadmon 2000, but see Casper 1996) and can even shift within a single year or  
54 season (Holzapfel and Mahall 1999, Kikvidze et al. 2006).

55           Herbivory also affects the net direction of plant-plant interactions by mediating both  
56 positive and negative interactions (Gurevitch et al. 2000, Hambäck and Beckerman 2003,

57 Callaway et al. 2005, Brooker et al. 2006, Graff et al. 2007). If a plant species is selectively  
58 grazed, its fitness may be decreased, thereby altering competitive dynamics within the plant  
59 community (Gurevitch et al. 2000, van der Wal et al. 2000, Fowler 2002). On the other hand,  
60 positive plant-plant relationships are mediated by herbivore foraging behavior when one plant  
61 species benefits by growing in close proximity to another (better-defended) species (e.g., Rebollo  
62 et al. 2002).

63         The behavior of herbivores and their effects on the plant community often vary  
64 seasonally. Mammalian herbivores, driven by metabolic requirements, often become more  
65 selective in order to attain high quality forage during dry times (Demment and Van Soest 1985,  
66 Belovsky 1997). Additionally, defoliation may be more intense or spatially concentrated as  
67 quantity and availability of forage becomes limited during dry times (Scoones 1995, Ilius and  
68 O'Connor 1999, Fynn and O'Connor 2000, Ilius and O'Connor 2000, Rousset and Lepar 2002).  
69 Conversely, defoliation may be less intense or more spatially diffuse during wet times when  
70 green forage is abundant (Scoones 1995, Ilius and O'Connor 2000).

71         Different combinations of herbivory and rainfall intensities have contrasting effects on  
72 community patterns and processes (Proulx and Mazumder 1998, Fynn and O'Connor 2000,  
73 Osem et al. 2002, Augustine and McNaughton 2006). Understanding how herbivory interacts  
74 with seasonal dynamics has direct implications for our understanding of the mechanics of  
75 competition-facilitation processes as drivers of short-term community change (e.g., Callaway et  
76 al. 2005). Additionally, short-term processes can influence the longer-term trajectory of the  
77 ecosystem (Hobbs 1996, Goldberg et al. 1999, Rousset and Lepar 2002, Gratton and Denno  
78 2003, Gómez 2005). Short-term competition-facilitation processes in particular can be key  
79 components of long-term successional change (Connell and Slatyer 1977, Olff et al. 1999,

80 Baumeister and Callaway 2006, Zanini et al. 2006, Dickie et al. 2007).

81 I experimentally examined the roles of season and herbivory in the competition-  
82 facilitation balance between two grass species, *Cynodon plectostachyus* and *Pennisetum*  
83 *stramineum*, in a semi-arid ecosystem. I manipulated neighboring plant composition and density,  
84 as well as herbivore access, in both wet and dry periods. These two grass species co-occur in  
85 large patches that undergo successional development from *Cynodon* dominance to *Pennisetum*  
86 dominance. I asked the following questions: a) Does facilitation occur between *Pennisetum* and  
87 *Cynodon*? b) If facilitation does occur, is it stronger in the dry season? c) Is facilitation  
88 mediated by herbivory? d) Do the effects of season and herbivory interact? and e) How do short-  
89 term dynamics compare to the known long-term trajectory of the system?

90

## 91 **Methods**

### 92 *Study site*

93 This research was conducted on the Mpala Research Centre and Farm and adjacent  
94 Segera and Jessel Ranches (36°52'E, 0°17'N), which cover more than 50,000 ha of semi-arid  
95 savanna in Laikipia, Kenya. Annual rainfall is variable, but averages approximately 500-600  
96 mm. The mean rainfall pattern is weakly trimodal, with the highest peak in mean rainfall  
97 occurring March-May (mean in April, 1998-2005: 120mm) and two lesser peaks in July and  
98 November (means 84 and 107, respectively). The longest and most pronounced dry season is  
99 typically from December through March.

100 My study area is located in *Acacia drepanolobium* savanna and is underlain with “black  
101 cotton” soils, poorly drained vertisols with high clay content (Ahn and Geiger 1987). In this  
102 system, 97% of the tree canopy cover is *Acacia drepanolobium*, and 90% of herbaceous cover is

103 comprised of six species: *Pennisetum mezianum*, *P. stramineum*, *Themeda triandra*, *Lintonia*  
104 *nutans*, *Aerva lanata* and *Brachiaria lachnantha* (Young et al. 1998). Resident large mammals  
105 include: elephants (*Loxodonta africana*), zebras (*Equus burchellii*), Grant's gazelles (*Gazella*  
106 *grantii*), hartebeests (*Alcelaphus buselaphus*), steinbucks (*Raphicerus campestris*), elands  
107 (*Taurotragus oryx*), buffaloes (*Syncerus caffer*), giraffes (*Giraffa camelopardalis*), spotted  
108 hyaenas (*Crocuta crocuta*), lions (*Panthera leo*), and leopards (*Panthera pardus*). This area is  
109 actively managed for cattle, with stocking rates of 0.5-0.7 livestock units/ha.

110         Historical and current cattle management has created a landscape mosaic of nutrient-rich,  
111 treeless "glades" scattered throughout the background of *A. drepanolobium* savanna. Once  
112 abandoned, temporary cattle corrals develop into these glades that are preferred by both livestock  
113 and wildlife. Vegetation in younger glades is dominated by the stoloniferous grass, *Cynodon*  
114 *plectostachyus*, while older glade vegetation is characterized by a near monoculture of an  
115 ascending bunchgrass, *Pennisetum stramineum* (K.Veblen, unpublished data). Glades of  
116 intermediate age are in transition and contain areas co-dominated by these two grass species.

117

#### 118 *Experimental design and data collection*

119         I investigated competition and facilitation dynamics between *Pennisetum stramineum* and  
120 *Cynodon plectostachyus* in glades using a 2 x 2 x 2 factorial design. For each of these two  
121 species, I crossed species removal treatments (interspecific neighbors present vs. interspecific  
122 neighbors removed) with grazing treatments (grazers allowed vs. grazers excluded).  
123 Experimental plots were randomly located within glade areas co-dominated by *Pennisetum* and  
124 *Cynodon* (plots were considered 'co-dominated' so long as neither species exhibited >80%  
125 cover; measurements described below).

126 Six neighbor/grazing treatment combinations were replicated within each of twelve  
127 randomly chosen glades (with treatments established in five glades in August 2004, and seven in  
128 August 2005) for a total of 72 experimental plots. Three of the six treatments were grazed, and  
129 the remaining three were ungrazed. Crossed with each grazing treatment were the following  
130 three species removal treatments: *Pennisetum* removal, *Cynodon* removal, and control (neither  
131 *Cynodon* nor *Pennisetum* removed). Each of the seven glades in 2005 included two additional  
132 treatments that were added to tease apart inter- vs. intraspecific neighbor effects: one grazed and  
133 one ungrazed density-reduced control plot. All 86 plots were 1x1m.

134 Grazed plots were left open to grazing, and ungrazed plots were covered with 1x1x1m  
135 chicken wire mesh cages that prevented grazing by larger vertebrate herbivores. The lower  
136 halves of the cages were covered with 30 cm tall hardware cloth to also exclude the dominant  
137 herbivorous rodent, *Saccostomus mearnsi*. In all plots, I identified all individual plants to  
138 species. In species removal plots, all individuals of target species were removed. In density-  
139 reduced control plots, I haphazardly removed individuals of both grass species until the overall  
140 cover was half of its starting density (with density first assessed visually and then confirmed by  
141 counting pin frame pin hits; see below). In all 86 plots, all forbs, which comprised an average of  
142 2.9% of total cover, and all non-target grass species, which comprised an average 2.8% of total  
143 cover, were removed. All removed plants were clipped to remove above-ground biomass and  
144 later were individually treated by sponge with Roundup® (glyphosate) herbicide after rainfall  
145 and any new growth.

146 Prior to manipulation, within each 1x1m plot, I quantified plant cover with five evenly  
147 spaced ten-point pin frames. I repeated baseline pin frame counts in 'removal' plots immediately  
148 after neighbor removal treatments were complete. I assessed the effects of experimental

149 treatments on *Cynodon* and *Pennisetum* by repeating pinframe measurements in all plots at the  
150 end of the three-month-long dry season in the March following plot set-up (28 mm rainfall in  
151 January-March 2005 and 24mm in January-March 2006). I repeated measurements three months  
152 later in June at the end of a heavy rainfall period (227 mm rainfall in April-June 2005, and 153  
153 mm in April-June 2006). Plots set up in 2004 were assessed only during the 2005 dry and wet  
154 seasons. For pin frame cover estimates, I counted the total number of times that each species  
155 touched each pin. As a result, the pin frame count for any given species often far exceeded ten.  
156 Although I hereafter refer to pin frame measurements as “cover,” total number of pin hits is  
157 positively correlated with biomass of each species (*Cynodon*  $R^2=0.89$ , K. Veblen unpublished  
158 data; *Pennisetum*  $R^2=0.65$ , I. Kimathi unpublished data).

159

#### 160 *Data analyses*

161 I calculated mean *Cynodon* and *Pennisetum* cover values for each 1x1m plot in each  
162 sampling period. I quantified how *Cynodon* and *Pennisetum* cover changed over time (in  
163 response to grazing, neighbor and season effects) by calculating a relative growth index.  
164 Relative growth was calculated as  $\text{Ln}((\text{cover } t_2)/(\text{cover } t_1))$ . I calculated relative growth during  
165 the dry season ( $t_1$ =post-manipulation baseline cover,  $t_2$ =March cover) and wet season ( $t_1$ =March  
166 cover,  $t_2$ =June cover) for each species/neighbor/grazing combination. Relative growth values are  
167 symmetrical around zero, with positive numbers indicating increases in cover and negative  
168 numbers indicating decreases in cover.

169 I calculated an additional index, relative interaction intensity (RII), to depict positive  
170 versus negative effects of heterospecific neighbor presence on cover of each species. The RII  
171 provides a measure of the relative difference in cover of a target species in plots with neighbors



172 versus plots without neighbors. I calculated the mean ( $\pm$  1 SE) relative interaction intensities  
173 (RII) between *Cynodon* and *Pennisetum* for all four season/grazing combinations (wet grazed,  
174 wet ungrazed, dry grazed and dry ungrazed). RII was calculated with the following equation:  
175  $RII = (B_w - B_o) / (B_w + B_o)$  where  $B_w$  = cover of target species growing with neighbor (i.e., control  
176 plot) and  $B_o$  = cover of target species growing without neighbors (i.e., removal plot) (Armas et al.  
177 2004). RII values fall between -1 and +1, with negative values indicating competitive neighbor  
178 effects and positive values indicating facilitative neighbor effects.

179 I used a general linear mixed model (GLMM) to conduct a repeated measures analysis of  
180 covariance (ANCOVA) testing treatment and time (i.e., season) effects on *Cynodon* and  
181 *Pennisetum* relative growth. For each grass species, grazing treatment (presence/absence),  
182 neighbor treatment (presence/absence), season (dry/wet), and interactions among the three were  
183 fixed factors. I used maximum-likelihood methodology, as well as Satterthwaite's  
184 approximation of degrees of freedom (PROC MIXED, version 9.1, SAS Institute 2002). Block  
185 (i.e., glade) was treated as the experimental unit and was designated as a random effect. Pre-  
186 manipulation baseline cover of target species was treated as the covariate. I made pairwise  
187 comparisons for grazing, competition and grazing\*competition within each season using  
188 unprotected LSD t-tests, applying Bonferonni corrections to significance values ( $p=0.025$ ,  $0.025$ ,  
189 and  $0.0125$ , respectively). For relative growth analyses, the two years' worth of data were  
190 pooled because species responses were qualitatively identical in the two years and preliminary  
191 models indicated that the effect of year was negligible. For density reduction plots, relative  
192 growth and statistical calculations were performed as above, replacing 'control' plots with  
193 'density-reduced control plots. Although mean RII values are presented to indicate the direction

194 and magnitude of neighbor effects, I did not conduct statistical analyses of RII values which are  
195 simply a composite representation of data already analyzed in relative growth analyses.

196

## 197 **Results**

198         The two grass species differed in their responses to experimental grazing and neighbor  
199 treatments over the two seasons. *Pennisetum* showed negative responses to grazing, especially  
200 during the dry season, and a highly positive response to rainfall. *Cynodon* also experienced  
201 strong competitive suppression under several treatment combinations. However, *Cynodon*  
202 showed a net facilitation response to neighboring *Pennisetum* plants in the presence of  
203 herbivores in the dry season.

204

### 205 *Pennisetum* performance

206         The presence of herbivores significantly decreased cover of *Pennisetum*, and this effect  
207 was especially pronounced during the dry season (Table 1; Fig. 1c,d). *Pennisetum* cover was  
208 44% lower in plots open to grazing. Although grazing decreased *Pennisetum* cover during the  
209 wet season (Fig. 1d), grazing led to net loss of *Pennisetum* cover only during the dry season (Fig.  
210 1c). Though less influential than grazing and non-significant, interspecific competitors also  
211 decreased *Pennisetum* (Fig. 1c; Table 1), with *Cynodon* removal increasing *Pennisetum* cover by  
212 an average of 16%. Although the season\*herbivory\*competitor interaction was not significant,  
213 (Table 1), the negative neighbor effects of *Cynodon* appear most pronounced in ungrazed plots  
214 during the dry season, (Fig. 1c). There was no evidence of facilitation of *Pennisetum* by  
215 *Cynodon* under any grazing by season combination (Fig. 1c,d; Table 1).

216 Despite the negative effects of grazing and neighbors on *Pennisetum* performance, wet  
217 season *Pennisetum* growth was so pronounced that net increases in cover occurred in all plots,  
218 even those with both competition and grazing. *Pennisetum* cover was 33% higher in the wet  
219 season than in the dry season. During the wet season *Pennisetum* was competitively dominant  
220 over *Cynodon*; *Pennisetum* cover increased (relative growth >0) in control plots (both with and  
221 without grazing), while *Cynodon* decreased (relative growth <0) (Fig. 1b,d).

222

### 223 *Cynodon* performance

224 *Cynodon* had a much more complex response to experimental treatments across sampling  
225 periods. During the dry season, grazing significantly reduced *Cynodon* cover ( $t=-2.96$ ,  $df=39.9$ ,  
226  $p=0.005$ ; Fig. 1a; Table 1). This effect was amplified by the removal of *Pennisetum*,  
227 demonstrating a facilitative effect ( $RII>0$ ) on *Cynodon* cover (Fig. 2a). *Cynodon* relative growth  
228 was significantly higher in the presence of its interspecific neighbor, but only in the presence of  
229 grazing (grazed, competitor pres/abs:  $t=3.19$ ,  $df=42.3$ ,  $p=0.0027$ ; ungrazed, competitor pres/abs:  
230  $t=0.74$ ,  $df=57.8$ ,  $p=0.33$ ; Table 1; Fig. 1a). However, this was not true during the wet season  
231 when *Pennisetum* had a strongly negative (competitive) effect on *Cynodon* cover ( $t=-4.49$   
232  $df=33.3$   $p<0.0001$ ; Table 1; Fig. 2a). Wet season grazing did not significantly reduce *Cynodon*  
233 cover overall ( $t=1.33$ ,  $df=39.9$ ,  $p=0.19$ ; Table 1; Fig. 1b), and in contrast to the dry season,  
234 *Cynodon* cover increased when subjected to both neighbor removal and grazing ( $t=-4.41$ ,  
235  $df=42.3$ ,  $p<0.0001$ ; Table 1; Fig. 1b). Net facilitation did not occur during the wet season.

236

### 237 Density reduction plots

238           The facilitative relationship between *Cynodon* and *Pennisetum* appears to be both  
239 unidirectional and species-specific. A dry season comparison of *Pennisetum* removal plots  
240 (leaving ~50% *Cynodon* cover) with density-reduced control plots (leaving ~25% *Pennisetum*  
241 and ~25% *Cynodon*) shows that, at a constant plant cover (~50%), *Cynodon* is facilitated only by  
242 *Pennisetum* (grazed, competitor pres/abs:  $t=3.76$ ,  $df=29.7$ ,  $p=0.0007$ ; ungrazed, competitor  
243 pres/abs:  $t=0.82$ ,  $df=33.1$ ,  $p=0.42$ ; Table 2; Fig. 3). This suggests that facilitation is due neither  
244 a) to less cover (regardless of species identity) or b) to presence of conspecifics.

245           *Cynodon* and *Pennisetum* performed as well or better in density-reduced plots than in  
246 neighbor removal plots in all season-grazing combinations except 1) *Cynodon*, grazed during dry  
247 season (discussed above) and 2) *Pennisetum*, ungrazed during wet season (data not presented).  
248 Overall, both species appeared to respond to a general reduction in competition (and increase in  
249 bare ground), irrespective of species identity. Additionally, *Pennisetum* appeared to respond  
250 (non-significantly) to higher inter- than intra-specific competition when ungrazed during the wet  
251 season.

252

## 253 **Discussion**

254           For a robust understanding of ecological interactions, ecologists must address context-  
255 dependency and study how biotic and abiotic factors vary over time to influence species'  
256 interactions (Agrawal et al. 2007). Semi-arid grazing ecosystems present an ideal opportunity to  
257 investigate how herbivory (biotic) effects and wet/dry (abiotic) cycles interact to influence plant-  
258 plant facilitation. This study provides experimental evidence that plant-plant interactions are  
259 highly contextual. Specifically, I have shown that facilitation 1) occurs only in the dry season, 2)  
260 occurs only in the presence of herbivores, 3) is asymmetric (*Cynodon* is facilitated by

261 *Pennisetum*, but not vice-versa), and 4) is species-specific (*Cynodon* benefits from the presence  
262 of *Pennisetum*, but not from the presence of con-specifics). In addition to revealing contextual  
263 facilitation of *Cynodon*, the design also revealed more general patterns of competitive (negative)  
264 interactions between the two species.

265

#### 266 Competitive dynamics

267 Both species showed strong responses to the wet season. *Pennisetum* responded to  
268 increased rainfall with clear competitive dominance and net growth, and *Cynodon* was strongly  
269 negatively affected by *Pennisetum* neighbors. These results are consistent with plant  
270 competition models indicating more intense competition at higher productivity (Grime 1973,  
271 1977, Tilman 1985, 1988), including during pulses of productivity in dry environments  
272 (Goldberg and Novoplansky 1997).

273 The sensitivity of *Pennisetum*, the competitive dominant, to grazing may help explain  
274 two aspects of my experimental results. First, it may partially explain why there were no clear  
275 competitive effects of *Cynodon* neighbors on *Pennisetum*; the strong sensitivity of *Pennisetum* to  
276 grazing would have swamped any sensitivity to competition. Second, the strong suppression of  
277 *Pennisetum* by grazing in the wet season may have ameliorated the effects of grazing on  
278 *Cynodon*; this would have contributed to the appearance of *Cynodon* as grazing tolerant under  
279 competition in the wet season (Figure 1b). Similarly, any negative effects of grazing on  
280 *Cynodon* also would have muted its competitive ability, thereby benefiting *Pennisetum*. In a  
281 Texas rangeland Fowler (2002) found that cattle grazing reduced plant-plant competition and  
282 attributed it to competitor damage, and a meta-analysis by Gurevitch et al. (2000) found evidence  
283 for a general pattern for predation (including herbivory) muting competition.

284

285 Dry season facilitation

286           Regardless of any herbivory tolerance, *Cynodon* was buffered against the intense  
287 herbivory of the dry season through a facilitative relationship with its *Pennisetum* competitors.  
288 In addition to being unidirectional and species specific (Fig. 1,4), facilitation was limited to the  
289 dry season only and occurred only in the presence of herbivores. This net facilitative effect was  
290 a product of increased strength of facilitation, but also may have been due to the decreased  
291 strength of competition during the dry season.

292           The predominance of facilitation during the dry season, but not the wet season, is  
293 consistent with the abiotic stress hypothesis and with studies showing that facilitation generally  
294 strengthens under drier conditions (Greenlee and Callaway 1996, Holzapfel and Mahall 1999,  
295 Pugnaire and Lazaro 2000, Tielborger and Kadmon 2000, Kikvidze et al. 2006). Facilitation in  
296 these cases is typically attributed to the positive effects of shading and its associated benefits  
297 (particularly reduction in transpiration rates) outweighing any negative (competitive) effects.  
298 My results also are consistent with recent suggestions that herbivory is the major driver of  
299 facilitation in water-limited environments (Baraza et al. 2006, Graff et al. 2007). In these  
300 systems, where water is a primary limiting resource (Noy-Meir 1973, Casper and Jackson 1997)  
301 dry conditions may amplify the negative (competitive) effects of neighbors for water and  
302 outweigh any positive neighbor effects (Tielborger and Kadmon 2000, Maestre et al. 2005).  
303 Instead, neighbors may mitigate another major stress: consumer (grazing) stress. My results  
304 demonstrate that neither abiotic stress alone, nor herbivory alone account for facilitation, but that  
305 facilitation is driven by an interaction between the two.

306           The contingency of facilitation upon both grazing and times of abiotic stress likely relates  
307 to grazing intensity. There is growing evidence pointing to the intensity (not simply presence or  
308 absence) of grazing as an important driver of facilitation (Rebollo et al. 2002, Graff et al. 2007),  
309 and the prevalence of facilitation is predicted to increase with increasing grazing stress (Bertness  
310 and Callaway 1994, but see Brooker et al. 2006, Smit et al. 2007). There are two major reasons  
311 to believe higher grazing intensity may have coincided with the dry season in my study: 1) dry  
312 seasons in general are associated with increases in grazing pressure on limited forage and/or with  
313 higher herbivore selectivity (Demment and Van Soest 1985, Scoones 1995, Belovsky 1997, Illius  
314 and O'Connor 1999, Fynn and O'Connor 2000, Illius and O'Connor 2000, Rousset and Lepar  
315 2002); and 2) in this particular study system, glade areas are subject to especially high herbivore  
316 use during dry periods (K.Veblen, unpublished data).

317

### 318 Associational resistance

319           Facilitation of *Cynodon* appears to be a case of “associational resistance” in which some  
320 plants gain an advantage by “hiding” among others of higher or lower palatability (Root 1973,  
321 Attsat and O'Dowd 1976, Hjältén et al. 1993, Callaway and Aschehoug 2000, Milchunas and  
322 Noy-Meir 2002, Bossuyt et al. 2005). Most examples of associational resistance relate to insect  
323 herbivores (e.g., Holmes and Jepson-Innes 1989, Hambäck et al. 2000). There are few explicit  
324 mammalian herbivore examples (Milchunas and Noy-Meir 2002, Hambäck and Beckerman  
325 2003), which more commonly fall into the category of shared mechanical defenses, such as  
326 spines (e.g., Callaway et al. 2000 and references therein). *Cynodon* is considered the most  
327 palatable common grass in this ecosystem (T. Young, personal communication; K. Veblen,  
328 interviews with resident ranchers and herders) and appears to gain associational resistance by

329 growing within less palatable *Pennisetum* stands. Aside from palatability, *Pennisetum* may  
330 further deter grazing in the dry season because its stalks and remaining shoots are thick and sharp  
331 after being grazed (K. Veblen, personal observation).

332 A major assumption of associational resistance is that herbivores are deterred by the less  
333 palatable neighbor species. The neighbor species must be of sufficiently low relative  
334 palatability. The grazer must then be sufficiently selective in the quality of food it seeks out  
335 because an extremely unselective grazer is more likely to graze all plants, regardless of  
336 palatability or quality (McNaughton 1978). A potential mechanism of associational resistance is  
337 that the grazer makes foraging decisions at the stand level (Hjältén et al. 1993), whereby the  
338 grazer decides to pass over a whole stand due to the unpalatability of only some plants within  
339 that stand. Because herbivores must optimize nutritional advantage (Arnold 1987, Milchunas  
340 and Noy-Meir 2002), herbivores may pass over a stand that includes less palatable *Pennisetum* if  
341 the costs of seeking out *Cynodon* outweigh nutritional benefits. Furthermore, *Cynodon*, which is  
342 characterized by long, thick, erect stolons that are difficult to visually distinguish from  
343 *Pennisetum* (K. Veblen, personal observation), may gain protection by being harder to detect  
344 when growing with *Pennisetum*.

345

#### 346 Long-term vegetation dynamics

347 Short-term species interactions have the potential to profoundly affect longer-term  
348 dynamics (Goldberg and Novoplansky 1997, Olff et al. 1999, Rousset and Lepart 2000, Armas  
349 and Pugnaire 2005). In this system, short-term competition-facilitation dynamics between  
350 *Cynodon* and *Pennisetum* may affect long-term glade successional patterns. Glade vegetation is  
351 initially colonized by *Cynodon* before gradual invasion and dominance by *Pennisetum* (K.



352 Veblen, unpublished data). This means that the wet season competitive dominance by  
353 *Pennisetum* may prevail over the long-term, which suggests that increased frequency or duration  
354 of wet periods could speed the succession process. Conversely, my results illustrate that dry  
355 season plant-plant interactions (such as the observed facilitation of *Cynodon* by *Pennisetum*)  
356 oppose the direction of succession, which may effectively retard or decrease the rate of glade  
357 succession.

358         Herbivory, including its intensity and interaction with wet/dry cycles, also is likely to  
359 mediate these successional processes. In savanna ecosystems, where the duration and intensity  
360 of herbivory and rainfall vary tremendously over time and space, herbivores—and their  
361 interactions with wet/dry cycles—have the potential to be major drivers of long-term ecosystem  
362 change. We know that herbivores can influence succession (e.g., Dormann et al. 2000, Fagan  
363 and Bishop 2000, Bokdam 2001, Skarpe 2001, Seabloom and Richards 2003), that intensity of  
364 herbivory matters (e.g., Rebollo et al. 2002, Brooker et al. 2006, Graff et al. 2007), and that  
365 different guilds of herbivores have different and sometimes interactive effects on the plant  
366 community (McNaughton 1978, Young et al. 2005). But we must explicitly test how herbivory  
367 interacts with temporal and spatial variation in plant community dynamics in order to elucidate  
368 and make meaningful predictions about the relationships between short-term and long-term  
369 dynamics in savanna ecosystems.

370

### 371 **Acknowledgments**

372 Many thanks to Frederick Erii, John Lochukuya, Jackson Ekadeli, and Patrick Etelej for their  
373 invaluable field assistance, great company and excellent eyesight. Thank you to Mpala Research  
374 Centre and staff, the Jessels, and Segera Ranch for logistical support and access to field sites.

375 Sarah Elmendorf , Kevin Rice, Corinna Riginos, Mark Schwartz and two anonymous reviewers  
376 provided helpful comments on earlier drafts of this paper. Statistical advice from Neil Willits is  
377 greatly appreciated. A special thanks to Truman Young for his stellar mentorship, advice,  
378 support, and readings of this manuscript at several stages. This research was funded by an NSF  
379 GRF, the UC Davis Biological Invasions IGERT, and NSF Grant BSR 03-16402.

380

### 381 **Literature Cited**

- 382 Agrawal, A. A., D. A. Ackerly, F. Adler, A. E. Arnold, C. Cáceres, D. F. Doak, E. Post, P. J.  
383 Hudson, J. Maron, K. A. Mooney, M. Power, D. Schemske, J. Stachowicz, S. Strauss, M.  
384 G. Turner, and E. Werner. 2007. Filling key gaps in population and community ecology.  
385 *Frontiers in ecology and the environment* **5**:145-152.
- 386 Ahn, P. M., and L. C. Geiger. 1987. Kenya Soil Survey-Soils of Laikipia District. Ministry of  
387 Agriculture, National Agricultural Laboratories, Kabete, Kenya.
- 388 Armas, C., R. Ordiales, and F. I. Pugnaire. 2004. Measuring plant interactions: A new  
389 comparative index. *Ecology* **85**:2682-2686.
- 390 Armas, C., and F. I. Pugnaire. 2005. Plant interactions govern population dynamics in a semi-  
391 arid plant community. *Journal of Ecology* **93**:978-989.
- 392 Arnold, G. W. 1987. Influence of the biomass, botanical composition and sward height of annual  
393 pastures on foraging behavior by sheep. *Journal of Applied Ecology* **24**:759-772.
- 394 Attsat, P. R., and D. J. O'Dowd. 1976. Plant defense guilds. *Science* **193**:24-29.
- 395 Augustine, D. J., and S. J. McNaughton. 2006. Interactive effects of ungulate herbivores, soil  
396 fertility, and variable rainfall on ecosystem processes in a semi-arid savanna. *Ecosystems*  
397 **9**:1242-1256.

- 398 Baraza, E., R. Zamora, and J. A. Hódar. 2006. Conditional outcomes in plant-herbivore  
399 interactions: neighbours matter. *Oikos* **113**:148-156.
- 400 Baumeister, D., and R. M. Callaway. 2006. Facilitation by *Pinus flexilis* during succession: A  
401 hierarchy of mechanisms benefits other plant species. *Ecology* **87**:1816-1830.
- 402 Belovsky, G. E. 1997. Optimal foraging and community structure: the allometry of herbivore  
403 food selection and competition. *Evolutionary Ecology* **11**:641-672.
- 404 Bertness, M. D., and R. M. Callaway. 1994. Positive interactions in communities. *Trends in*  
405 *Ecology and Evolution* **9**:191-193.
- 406 Bokdam, J. 2001. Effects of browsing and grazing on cyclic succession in nutrient-limited  
407 ecosystems. *Journal of Vegetation Science* **12**:875-886.
- 408 Bossuyt, B., B. De Fre, and M. Hoffmann. 2005. Abundance and flowering success patterns in a  
409 short-term grazed grassland: early evidence of facilitation. *Journal of Ecology* **93**:1104-  
410 1114.
- 411 Brooker, R. W., D. Scott, S. C. F. Palmer, and E. Swaine. 2006. Transient facilitative effects of  
412 heather on Scots pine along a grazing disturbance gradient in Scottish moorland. *Journal*  
413 *of Ecology* **94**:637-645.
- 414 Callaway, R. M. 1992. Effect of shrubs on recruitment of *Quercus douglasii* and *Quercus lobata*  
415 in California. *Ecology* **73**:2118-2128.
- 416 Callaway, R. M. 1995. Positive interactions among plants. *The Botanical Review* **61**:306-349.
- 417 Callaway, R. M., and E. T. Aschehoug. 2000. Invasive plants versus their new and old  
418 neighbors: A mechanism for exotic invasion. *Science* **290**:521-523.

- 419 Callaway, R. M., R. W. Brooker, P. Choler, Z. Kikvidze, C. J. Lortie, R. Michalet, L. Paolini, F.  
420 I. Pugnaire, B. Newingham, E. T. Aschehoug, C. Armas, D. Kikodze, and B. J. Cook.  
421 2002. Positive interactions among alpine plants increase with stress. *Nature* **417**:844-848.
- 422 Callaway, R. M., D. Kikodze, M. Chiboshvili, and L. Khetsuriani. 2005. Unpalatable plants  
423 protect neighbors from grazing and increase plant community diversity. *Ecology*  
424 **86**:1856-1862.
- 425 Callaway, R. M., Z. Kikvidze, and D. Kikodze. 2000. Facilitation by unpalatable weeds may  
426 conserve plant diversity in overgrazed meadows in the Caucasus Mountains. *Oikos*  
427 **89**:275-282.
- 428 Casper, B. B., and R. B. Jackson. 1997. Plant competition underground. *Annual Review of*  
429 *Ecology and Systematics* **28**:545-570.
- 430 Connell, J. H. 1983. On the prevalence and relative importance of interspecific competition -  
431 evidence from field experiments. *American Naturalist* **122**:661-696.
- 432 Connell, J. H., and R. O. Slatyer. 1977. Mechanisms of succession in natural communities and  
433 their role in community stability and organization. *American Naturalist* **111**:1119-1144.
- 434 Demment, M. W., and P. J. Van Soest. 1985. A nutritional explanation for body size patterns of  
435 ruminant and nonruminant herbivores. *American Naturalist* **125**:641-672.
- 436 Dickie, I. A., S. A. Schnitzer, P. B. Reich, and S. E. Hobbie. 2007. Is oak establishment in old-  
437 fields and savanna openings context dependent? *Journal of Ecology* **95**:309-320.
- 438 Dormann, C. F., R. Van Der Wal, and J. P. Bakker. 2000. Competition and herbivory during salt  
439 marsh succession: The importance of forb growth strategy. *Journal of Ecology* **88**:571-  
440 583.

- 441 Fagan, W. F., and J. G. Bishop. 2000. Trophic interactions during primary succession: herbivores  
442 slow a plant reinvasion at Mount St. Helens. *American Naturalist* **155**:238-251.
- 443 Fowler, N. L. 2002. The joint effects of grazing, competition, and topographic position on six  
444 savanna grasses. *Ecology* **83**:2477-2488.
- 445 Fynn, R. W. S., and T. G. O'Connor. 2000. Effect of stocking rate and rainfall on rangeland  
446 dynamics and cattle performance in a semi-arid savanna, South Africa. *Journal of*  
447 *Applied Ecology* **37**:491-507.
- 448 Goldberg, D., and A. Novoplansky. 1997. On the relative importance of competition in  
449 unproductive environments. *Journal of Ecology* **85**:409-418.
- 450 Goldberg, D. E., and A. M. Barton. 1992. Patterns and consequences of interspecific competition  
451 in natural communities: a review of field experiments with plants. *American Naturalist*  
452 **139**:771-801.
- 453 Goldberg, D. E., T. Rajaniemi, J. Gurevitch, and A. Stewart-Oaten. 1999. Empirical approaches  
454 to quantifying interaction intensity: Competition and facilitation along productivity  
455 gradients. *Ecology* **80**:1118-1131.
- 456 Gómez, J. M. 2005. Long-term effects of ungulates on performance, abundance, and spatial  
457 distribution of two montane herbs. *Ecological Monographs* **75**:231-258.
- 458 Gómez, J. M., J. A. Hódar, R. Zamora, J. Castro, and D. García. 2001. Ungulate damage on  
459 Scots pines in Mediterranean environments: effects of association with shrubs. *Canadian*  
460 *Journal of Botany-Revue Canadienne De Botanique* **79**:739-746.
- 461 Graff, P., M. R. Auguiar, and E. J. Chaneton. 2007. Shifts in positive and negative plant  
462 interactions along a grazing intensity gradients. *Ecology* **88**:188-199.

- 463 Gratton, C., and R. F. Denno. 2003. Inter-year carryover effects of a nutrient pulse on *Spartina*  
464 plants, herbivores, and natural enemies. *Ecology* **84**:2692-2707.
- 465 Greenlee, J. T., and R. M. Callaway. 1996. Abiotic stress and the relative importance of  
466 interference and facilitation in montane bunchgrass communities in western Montana.  
467 *American Naturalist* **148**:386-396.
- 468 Grime, J. P. 1973. Competition and diversity in herbaceous vegetation. *Nature* **244**:310-311.
- 469 Grime, J. P. 1977. Evidence for the existence of three primary strategies in plants and its  
470 relevance to ecological and evolutionary theory. *The American Naturalist* **111**:1169-  
471 1194.
- 472 Gurevitch, J., J. A. Morrison, and L. V. Hedges. 2000. The interaction between competition and  
473 predation: a meta-analysis of field experiments. *The American Naturalist* **155**:435-453.
- 474 Hambäck, P. A., J. Agren, and L. Ericson. 2000. Associational resistance: Insect damage to  
475 purple loosestrife reduced in thickets of sweet gale. *Ecology* **81**:1784-1794.
- 476 Hambäck, P. A., and A. P. Beckerman. 2003. Herbivory and plant resource competition: a  
477 review of two interacting interactions. *Oikos* **101**:26-37.
- 478 Hjältén, J., K. Danell, and P. Lundberg. 1993. Herbivore avoidance by association - vole and  
479 hare utilization of woody plants. *Oikos* **68**:125-131.
- 480 Hobbs, N. T. 1996. Modification of ecosystems by ungulates. *Journal of Wildlife Management*  
481 **60**:695-713.
- 482 Holmes, R. D., and K. Jepson-Innes. 1989. A neighborhood analysis of herbivory in *Bouteloua*  
483 *gracilis*. *Ecology* **70**:971-976.
- 484 Holmgren, M., M. Scheffer, and M. A. Huston. 1997. The interplay of facilitation and  
485 competition in plant communities. *Ecology* **78**:1966-1975.

- 486 Holzapfel, C., and B. E. Mahall. 1999. Bidirectional facilitation and interference between shrubs  
487 and annuals in the Mojave Desert. *Ecology* **80**:1747-1761.
- 488 Illius, A. W., and T. G. O'Connor. 2000. Resource heterogeneity and ungulate population  
489 dynamics. *Oikos* **89**:283-294.
- 490 Illius, A. W., and T. G. O'Connor. 1999. On the relevance of nonequilibrium concepts to arid and  
491 semiarid grazing systems. *Ecological Applications* **9**:798-813.
- 492 Kikvidze, Z., L. Khetsuriani, D. Kikodze, and R. M. Callaway. 2006. Seasonal shifts in  
493 competition and facilitation in subalpine plant communities of the central Caucasus.  
494 *Journal of Vegetation Science* **17**:77-82.
- 495 Maestre, F. T., S. Bautista, and J. Cortina. 2003. Positive, negative, and net effects in grass-shrub  
496 interactions in mediterranean semiarid grasslands. *Ecology* **84**:3186-3197.
- 497 Maestre, F. T., F. Valladares, and J. F. Reynolds. 2005. Is the change of plant-plant interactions  
498 with abiotic stress predictable? A meta-analysis of field results in arid environments.  
499 *Journal of Ecology* **93**:748-757.
- 500 McNaughton, S. J. 1978. Serengeti ungulates: feeding selectivity influences the effectiveness of  
501 plant defense guilds. *Science* **199**:806-807.
- 502 Milchunas, D. G., and I. Noy-Meir. 2002. Grazing refuges, external avoidance of herbivory and  
503 plant diversity. *Oikos* **99**:113-130.
- 504 Noy-Meir, I. 1973. Desert ecosystems: environment and producers. *Annual Review of Ecology  
505 & Systematics* **4**:25-51.
- 506 Olf, H., F. W. M. Vera, J. Bokdam, E. S. Bakker, J. M. Gleichman, K. de Maeyer, and R. Smit.  
507 1999. Shifting mosaics in grazed woodlands driven by the alternation of plant facilitation  
508 and competition. *Plant Biology* **1**:127-137.

- 509 Osem, Y., A. Perevolotsky, and J. Kigel. 2002. Grazing effect on diversity of annual plant  
510 communities in a semi-arid rangeland: interactions with small-scale spatial and temporal  
511 variation in primary productivity. *Journal of Ecology* **90**:936-946.
- 512 Proulx, M., and A. Mazumder. 1998. Reversal of grazing impact on plant species richness in  
513 nutrient-poor vs. nutrient-rich ecosystems. *Ecology* **79**:2581-2592.
- 514 Pugnaire, F. I., and R. Lazaro. 2000. Seed bank and understorey species composition in a semi-  
515 arid environment: The effect of shrub age and rainfall. *Annals of Botany* **86**:807-813.
- 516 Pugnaire, F. I., and M. T. Luque. 2001. Changes in plant interactions along a gradient of  
517 environmental stress. *Oikos* **93**:42-49.
- 518 Rebollo, S., D. G. Milchunas, I. Noy-Meir, and P. L. Chapman. 2002. The role of a spiny plant  
519 refuge in structuring grazed shortgrass steppe plant communities. *Oikos* **98**:53-64.
- 520 Root, R. B. 1973. Organization of a plant-arthropod association in simple and diverse habitats:  
521 the fauna of collards (*Brassica oleracea*). *Ecological Monographs* **43**:95-124.
- 522 Rousset, O., and J. Lepart. 2000. Positive and negative interactions at different life stages of a  
523 colonizing species (*Quercus humilis*). *Journal of Ecology* **88**:401-412.
- 524 Rousset, O., and J. Lepart. 2002. Neighbourhood effects on the risk of an unpalatable plant being  
525 grazed. *Plant Ecology* **165**:197-206.
- 526 Schoener, T. W. 1983. Field experiments on interspecific competition. *American Naturalist*  
527 **122**:240-285.
- 528 Scoones, I. 1995. Exploiting heterogeneity: habitat use by cattle in dryland Zimbabwe. *Journal of*  
529 *Arid Environments* **29**:221-237.
- 530 Seabloom, E. W., and S. A. Richards. 2003. Multiple stable equilibria in grasslands mediated by  
531 herbivore population dynamics and foraging behavior. *Ecology* **84**:2891-2904.



- 532 Skarpe, C. 2001. Effects of large herbivores on competition and succession in natural savannah  
533 rangelands. Pages 175-191. *in* P. Tow and A. Lazenby, editors. Competition and  
534 succession in pastures. CABI Publishing, New York.
- 535 Smit, C., C. Vandenberghe, J. den Ouden, and H. Müller-Schärer. 2007. Nurse plants, tree  
536 saplings and grazing pressure: changes in facilitation along a biotic environmental  
537 gradient. *Oecologia* **152**:265-273.
- 538 Tielborger, K., and R. Kadmon. 2000. Temporal environmental variation tips the balance  
539 between facilitation and interference in desert plants. *Ecology* **81**:1544-1553.
- 540 Tilman, D. 1985. The resource-ratio hypothesis of plant succession. *American Naturalist*  
541 **125**:827-852.
- 542 Tilman, D. 1988. Plant strategies and the dynamics and structure of plant communities. Princeton  
543 University Press, Princeton, New Jersey, USA.
- 544 van der Wal, R., M. Egas, A. van der Veen, and J. Bakker. 2000. Effects of resource competition  
545 and herbivory on plant performance along a natural productivity gradient. *Journal of*  
546 *Ecology* **88**:317-330.
- 547 Young, T. P., B. D. Okello, D. Kinyua, and T. M. Palmer. 1998. KLEE: a long-term multi-  
548 species herbivore exclusion experiment in Laikipia, Kenya. *African Journal of Range and*  
549 *Forage Science* **14**:94-102.
- 550 Young, T. P., T. M. Palmer, and M. E. Gadd. 2005. Competition and compensation among cattle,  
551 zebras, and elephants in a semi-arid savanna in Laikipia, Kenya. *Biological conservation*  
552 **122**:351-359.
- 553 Zanini, L., G. Ganade, and I. Hübel. 2006. Facilitation and competition influence succession in a  
554 subtropical old field. *Plant Ecology* **185**:179-190.

555 Table 1. Effects of grazing (presence/absence), competitors (presence/absence) and season (dry/wet)  
 556 on relative growth of *Cynodon plectostachyus* and *Pennisetum stramineum* in 1x1m plots.

557

Factor	<i>Cynodon plectostachyus</i>			<i>Pennisetum stramineum</i>		
	df	F	p	df	F	p
Grazing pres/abs	1, 22.0	23.34	<.0001	1, 20.8	30.65	<.0001
Competitor pres/abs	1, 23.4	6.57	0.017	1, 43.0	3.26	0.078
Grazing*Competitor	1, 21.8	3.53	0.074	1, 44.0	0.16	0.69
Season	1, 21.8	0.59	0.45	1, 21.8	3.29	0.083
Season*Grazing	1, 22.0	34.67	<.0001	1, 20.8	5.05	0.034
Season*Competitor	1, 21.8	27.21	<.0001	1, 43.0	0.85	0.36
Season*Grazing*Competitor	1, 21.8	4.74	0.04	1, 43.0	2.35	0.13
Baseline covariate	1, 67.7	3.52	0.065	1, 82.9	8.79	0.0039

558

559 Notes: Each species was analyzed as a general linear mixed model (GLMM) repeated measures  
 560 ANCOVA. Grazing, competitors, season, and interactions among the three were fixed effects. The  
 561 experimental unit, block, was a random effect, and baseline plant cover of the target species was  
 562 treated as the covariate.

563 Table 2. Effects of grazing (presence/absence), competitors (presence/absence) and season (dry/wet)  
 564 on relative growth of *Cynodon plectostachyus* and *Pennisetum stramineum* in density-reduced 1x1m  
 565 plots. In plots where competitors were present, overall density of both species was reduced to ~50%  
 566 to match overall density of competitor absence plots.

567

<i>Density-reduced</i>						
Factor	<i>Cynodon plectostachyus</i>			<i>Pennisetum stramineum</i>		
	df	F	p	df	F	p
Grazing pres/abs	1, 23.1	24.09	<.0001	1, 18.1	28.08	<.0001
Competitor pres/abs	1, 22.8	0.82	0.37	1, 13.7	0.50	0.49
Grazing*Competitor	1, 20.8	1.32	0.26	1, 9.48	0.27	0.62
Season	1, 24.7	1.09	0.31	1, 22.2	4.64	0.042
Season*Grazing	1, 22.4	22.11	0.0001	1, 17.8	6.58	0.02
Season*Competitor	1, 17.9	8.03	0.011	1, 13.6	2.60	0.13
Season*Grazing*Competitor	1, 18.7	8.90	0.0078	1, 9.38	0.41	0.54
Baseline covariate	1, 58.0	1.22	0.27	1, 42.6	8.39	0.0059

568

569 Notes: Each species was analyzed as a general linear mixed model (GLMM) repeated measures  
 570 ANCOVA. Grazing, competitors, season, and interactions among the three were fixed effects. Block  
 571 was a random effect, and baseline plant cover of the target species was treated as the covariate.

572

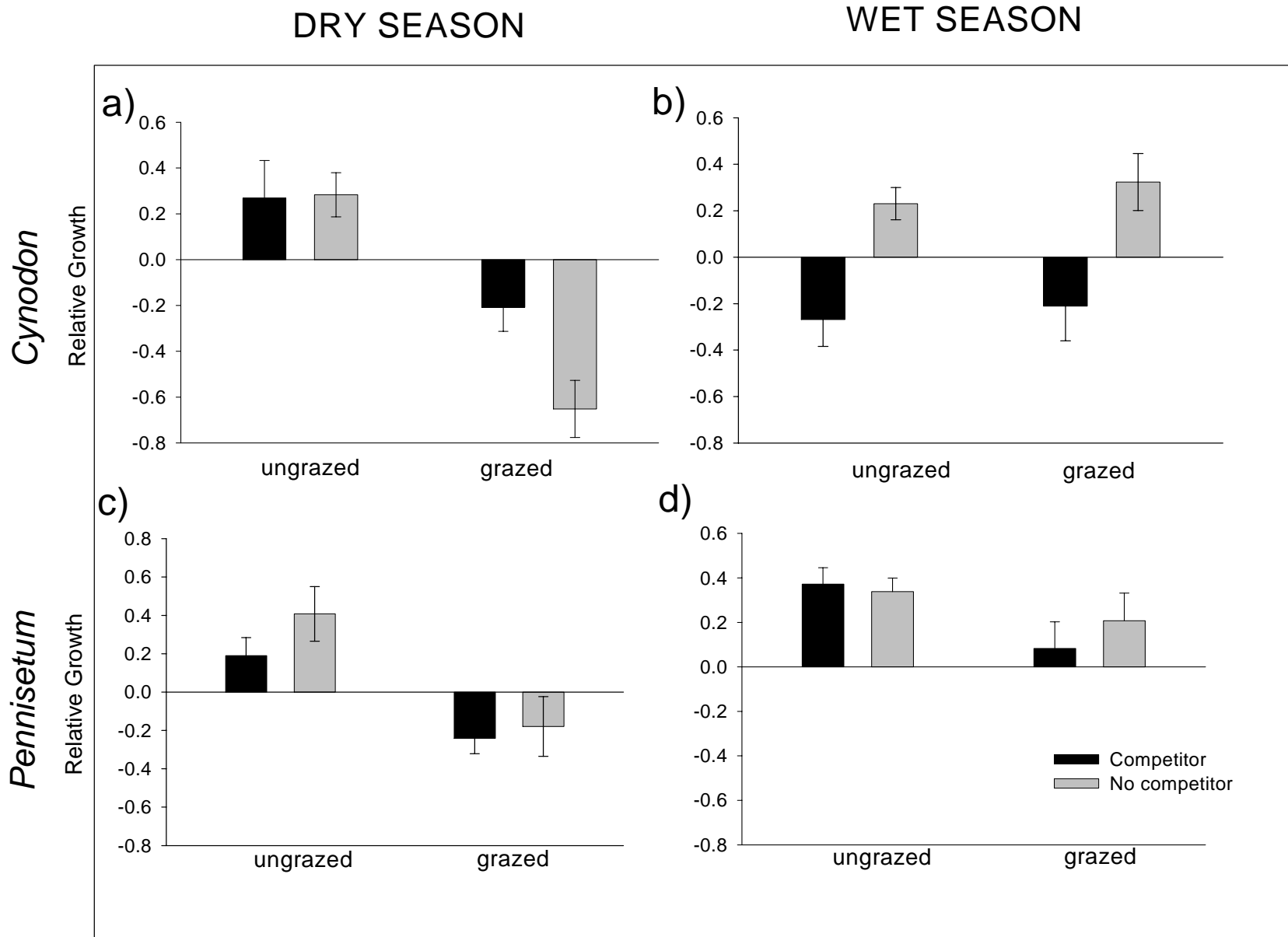
573 **Figure legends**

574 Figure 1. Effects of interspecific competitors (presence/absence) and grazing (presence/absence)  
575 on mean ( $\pm$ 1SE) relative growth of *Cynodon plectostachyus* (a, b) and *Pennisetum stramineum*  
576 (c,d) in 1x1m. plots during dry and wet seasons. Relative growth calculated as  $\text{Ln}[(\text{cover}$   
577  $t_2)/(\text{cover } t_1)]$ .

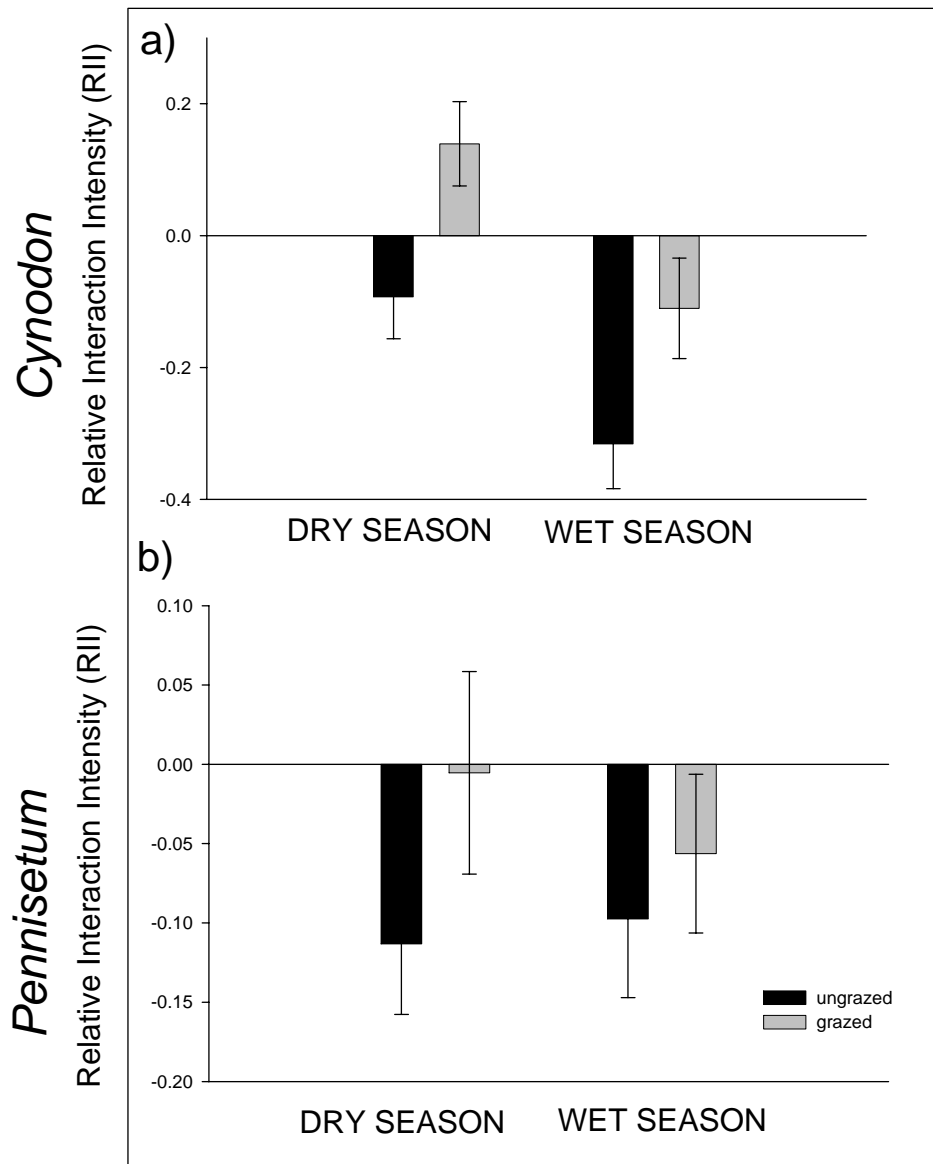
578  
579 Figure 2. Effects of season (dry/wet) and grazing (presence/absence) on mean ( $\pm$ 1SE) relative  
580 interaction intensity (RII) between *Cynodon plectostachyus* and *Pennisetum stramineum* in 1x1m  
581 plots. RII ranges from -1 to 1, with positive numbers indicating a positive (facilitative) effect of  
582 interspecific neighbors and negative numbers indicating a negative (competitive) effect of neighbors.  
583 The effects of *Pennisetum* on *Cynodon* are indicated in a), and the reverse in b). Note different scales  
584 of y axes.

585  
586 Figure 3. Effects of grazing (presence vs. absence) and competitor removal (inter- and intra-specific  
587 vs. only interspecific) on mean ( $\pm$ 1SE) relative growth of *Cynodon plectostachyus* during the dry  
588 season in 1x1m plots. In *Pennisetum* removal plots, all *Pennisetum* plants were removed. In  
589 *Cynodon+Pennisetum* removal plots, half of each species was removed. Overall density of plants was  
590 held constant across all plots.

591 **Figure 1.**

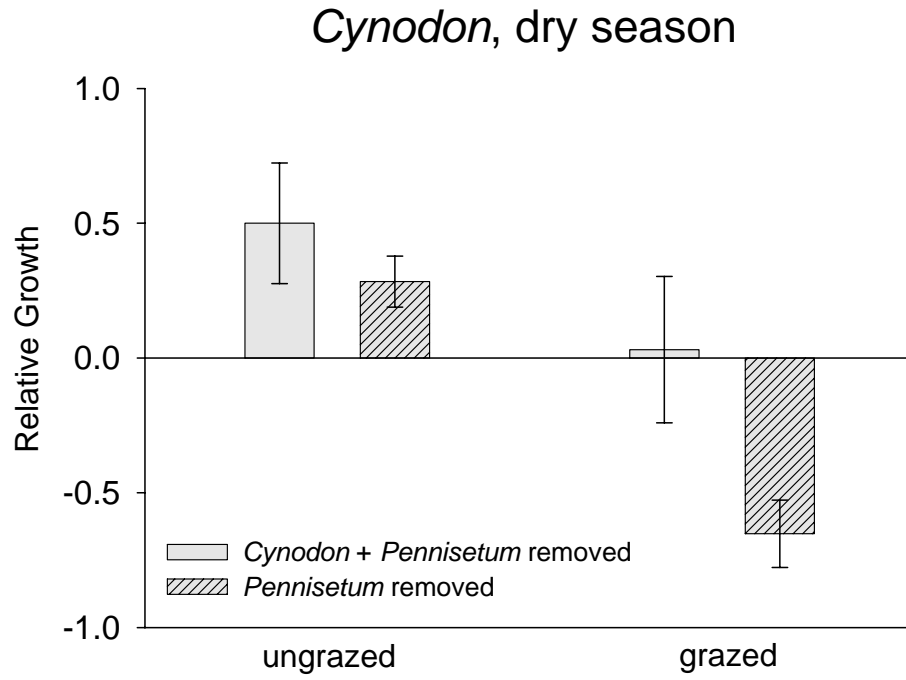


593 **Figure 2.**  
594



595  
596

597 **Figure 3.**



598